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Réunion des points focaux du MED POL

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Point 12 de l'ordre du jour : Harmonisation et normalisation de la surveillance du cluster IMAP Pollution

a) Directives / protocoles de suivi pour les indicateurs communs IMAP 13, 14, 17, 18, 20 et 23

b) Directives / protocoles de surveillance pour l'assurance qualité analytique et la communication des données de surveillance pour les indicateurs communs IMAP 13, 14, 17, 18 et 20

c) Directives / protocoles de surveillance pour les microplastiques flottants

Directives/Protocoles de contrôle concernant le prélèvement et la conservation des échantillons d'eau de mer pour l'indicateur commun 17 de l'IMAP : métaux lourds et éléments traces et contaminants organiques

PNUE/PAM
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Note du Secrétariat

Conformément au programme de travail 2020-2021 adopté par la COP21, le programme MED POL a préparé les lignes directrices de surveillance relatives aux indicateurs communs 13, 14, 17 et 20 de l'IMAP en vue de leur examen lors de la réunion intégrée des groupes de correspondance sur la surveillance de l'approche écosystémique (CORMON) (décembre 2020), tandis que les lignes directrices de surveillance pour l'indicateur commun 18 ainsi que les lignes directrices de surveillance relatives à l'assurance qualité et à la communication des données sont en cours de finalisation en vue de leur examen lors de la réunion du CORMON sur la surveillance de la pollution prévue en avril 2021.

Ces lignes directrices de surveillance contiennent des manuels cohérents destinés à guider le personnel technique des laboratoires compétents IMAP des Parties contractantes pour la mise en œuvre des pratiques de surveillance normalisées et harmonisées liées à un indicateur commun IMAP spécifique (c'est-à-dire l'échantillonnage, la conservation et le transport des échantillons, la préparation et l'analyse des échantillons, ainsi que l'assurance qualité et la communication des données de surveillance). Pour la première fois, ces lignes directrices présentent un résumé des meilleures pratiques connues disponibles et utilisées dans la surveillance du milieu marin, en exposant des pratiques analytiques globales intégrées qui pourront être appliquées afin de garantir la représentativité et l'exactitude des résultats analytiques nécessaires à la production de données de surveillance de qualité assurée.

Les lignes directrices/protocoles de surveillance s'appuient sur les connaissances et les pratiques acquises au cours des 40 années de mise en œuvre de la surveillance du MED POL et sur des publications récentes, mettant en évidence les pratiques actuelles des laboratoires maritimes des Parties contractantes ainsi que d'autres pratiques issues des conventions sur les mers régionales et de l'Union européenne. Une analyse approfondie des pratiques actuellement disponibles du PNUE/PAM, du PNUE et de l'AIEA ainsi que d'HELCOM, d'OSPAR et du Centre commun de recherche de la Commission européenne a été entreprise afin de contribuer à une approche novatrice pour la préparation des lignes directrices/protocoles de surveillance de l'IMAP.
Les lignes directrices/protocoles de surveillance abordent également les problèmes identifiés lors de la réalisation des épreuves de compétence organisées par l'UNEP/MAP-MEDPOL et l'AIEA depuis deux décennies maintenant, les nombreux résultats insatisfaisants dans le cadre des tests inter laboratoires pouvant être liés à des pratiques inadéquates au sein des laboratoires compétents de l'IMAP/MEDPOL.

L'eau de mer n'est pas incluse dans les matrices obligatoires à analyser dans le cadre du programme intégré de surveillance et d'évaluation (IMAP) de l'UNEP/MAP. La mise en œuvre d'un programme de surveillance pour déterminer la concentration en métaux lourds et en contaminants organiques dans l'eau de mer est donc une décision qui relève des pays. Afin de soutenir les efforts nationaux, les présentes lignes directrices de surveillance fournissent une note technique sur l'échantillonnage et le prétraitement de l'eau de mer en vue de l'analyse des métaux lourds et des contaminants organiques, comprenant les six protocoles suivants : i) Protocole d'échantillonnage de l'eau de mer pour l'analyse des métaux lourds ; ii) Protocole de filtration de l'eau de mer (métaux lourds) ; iii) Protocole de stockage à bord d'échantillons d'eau de mer pour l'analyse des métaux lourds ; iv) Protocole d'échantillonnage de l'eau de mer pour l'analyse des contaminants organiques ; v) Protocole de filtration de l'eau de mer (contaminants organiques) ; et vi) Protocole de stockage à bord d'échantillons d'eau de mer pour l'analyse des contaminants organiques.

Les lignes directrices/protocoles de surveillance, y compris celui-ci relatif à l'échantillonnage et à la préservation des échantillons d'eau de mer pour l'analyse de l'indicateur commun 17 de l'IMAP, établissent une base solide pour une mise à jour régulière des pratiques de surveillance en vue d'une mise en œuvre réussie de l'IMAP.

Conformément aux conclusions et recommandations des réunions intégrées des groupes de correspondance sur la mise en œuvre de l'approche écosystémique de l'IMAP (CORMON)

(vidéoconférence, 1-3 décembre 2020), et en particulier au paragraphe 22, la Réunion des CORMON a demandé au Secrétariat de modifier les Lignes directrices/Protocoles de surveillance en abordant les propositions techniques convenues qui ont été décrites dans le rapport de la Réunion et de soumettre l'ensemble de ces documents à la réunion des points focaux du MED POL. Les amendements demandés comprenaient des suggestions techniques écrites qui ont été fournies par plusieurs Parties contractantes jusqu'à 10 jours après la réunion intégrée des CORMON. Le document amendé a été partagé par le Secrétariat le 19 février 2021 pour une période de 2 semaines pour la non-objection des réunions intégrées des CORMON sur les changements introduits. Suite à l'absence d'objection de la réunion intégrée des CORMON, cette directive de suivi est soumise à l'examen de la présente réunion des points focaux MED POL.

Liste des abréviations / acronymes

CI	Indicateur commun
CdP	Conférence des parties
CORMON	Groupe de correspondance sur la surveillance
EcAp	Approche écosystémique
EEA	Agence environnementale européenne
EC	Commission européenne
EU	Union européenne
FAO	Organisation des Nations Unies pour l'alimentation et l'agriculture
GEOTRACES	Une étude internationale des cycles biogéochimiques marins des oligo-éléments et des isotopes
HELCOM	Commission pour la protection du milieu marin dans la zone de la mer Baltique – Commission d'Helsinki
HPDE	Polyéthylène haute densité
IAEA	Agence internationale de l'énergie atomique
IOC	Commission océanographique intergouvernementale
IMAP	Programme de surveillance et d'évaluation intégrées de la mer et des côtes méditerranéennes et les critères d'évaluation connexes
MAP	Plan d'action pour la Méditerranée
MED POL	Programme coordonné de surveillance continue et de recherche en matière de pollution dans la Méditerranée
MED QSR	Rapport sur la qualité de la Méditerranée
OSPAR	Convention pour la protection du milieu marin de l'Atlantique du nord-est
PEBD	Polyéthylène basse densité
PoW	Programme de travail
QA/QC	Assurance qualité / Contrôle qualité
QSR	Rapport sur la qualité

1 Introduction

1. L'eau de mer n'est pas incluse dans les matrices obligatoires à analyser dans le cadre du programme intégré de surveillance et d'évaluation de l'UNEP/MAP (UNEP 2019a1,UNEP 2019b2), par conséquent la mise en œuvre d'un programme de surveillance pour déterminer la concentration des métaux lourds et des contaminants organiques dans l'eau de mer est une décision qui relève des pays. Il faut souligner que les concentrations en métaux lourds et en contaminants organiques dans l'eau de mer sont très faibles, en particulier dans les eaux au large, de sorte qu'un prélèvement et une manipulation incorrects des échantillons peuvent facilement entraîner une perte de déterminant et/ou une contamination de l'échantillon avant l'analyse. Par conséquent, si un pays décide de mettre en œuvre un programme de surveillance de l'eau de mer, il doit élaborer et tester un protocole d'échantillonnage et de préservation très strict, en utilisant un équipement et une infrastructure de transport appropriés. De plus, les installations de laboratoire doivent être adaptées en conséquence pour assurer la qualité des analyses de concentrations en contaminants ultra-faibles dans les échantillons d'eau de mer.

2. L'échantillonnage de l'eau de mer peut être mis en œuvre de la même manière dans les zones marines côtières et offshore, puisque les équipements d'échantillonnage et les méthodes de préservation des échantillons pour éviter la perte de déterminants et/ou la contamination croisée sont similaires pour les deux zones. Par conséquent, les protocoles d'échantillonnage et de conservation suggérés sont applicables aussi bien dans les stations d'échantillonnage côtières que dans les stations offshores, en tenant compte du fait que les concentrations en métaux lourds et de contaminants organiques dans les eaux offshore sont susceptibles d'être plus faibles que dans les échantillons d'eau de mer côtière. Si des transects sont échantillonnés, l'échantillonnage doit se faire du large vers la côte et non l'inverse pour éviter la contamination des échantillons par le matériel d'échantillonnage.

3. Il est important de prélever des échantillons d'eau de mer représentatifs dans la zone d'échantillonnage, mais il est tout aussi important d'éviter toute altération des caractéristiques physiques et chimiques des échantillons pendant leur transport du lieu de prélèvement vers le laboratoire. Par conséquent, le stockage et le transport de l'eau de mer doivent être effectués selon des procédures spécifiques, afin d'éviter l'altération des échantillons et la contamination croisée par les matériaux des contenants et l'environnement de transport.

4. Pour aider les pays qui prévoient d'inclure la surveillance de l'eau de mer dans leurs programmes nationaux respectifs de surveillance de l'IC17, en tant que décision nationale, des protocoles d'échantillonnage et de traitement des échantillons d'eau de mer ont été élaborés. Ces protocoles sont conçus pour être non pas des manuels de formation analytique, mais des lignes directrices pour les laboratoires méditerranéens, qu'il convient de tester et de modifier afin d'en valider les résultats finaux.

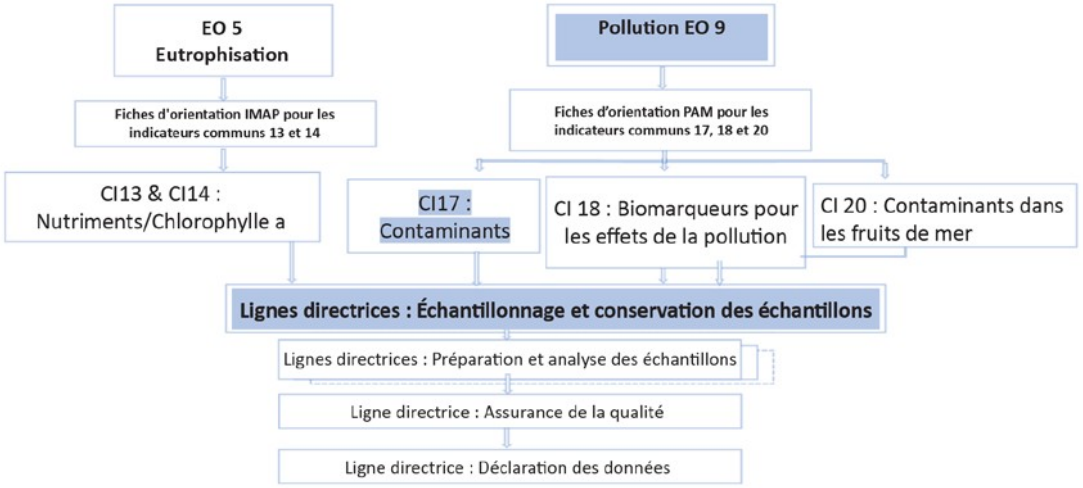
5. Les protocoles visent à rationaliser l'échantillonnage et le traitement des échantillons d'eau de mer en vue de garantir une assurance qualité comparable des données, ainsi que la comparabilité entre les zones d'échantillonnage et les différents programmes de surveillance nationaux. Ils fournissent des conseils, étape par étape, sur les méthodes à appliquer dans la zone méditerranéenne en matière d'échantillonnage, de manipulation des échantillons pour éviter la contamination croisée, ainsi que sur les conditions de stockage en vue de maintenir l'intégrité de l'échantillon pendant le transfert du site d'échantillonnage vers le laboratoire d'analyse et de garantir la représentativité et l'intégrité des échantillons à analyser.

6. En vue d'éviter les répétitions inutiles, il est également fait référence aux protocoles déjà publiés et accessibles au public, qui peuvent également être utilisés par les laboratoires compétents des Parties contractantes participant à la mise en œuvre de l'IMAP. Les six protocoles IMAP élaborés ci-dessous s'appuient sur les lignes directrices pertinentes élaborées par GEOTRACES, CIEM/OSPAR et HELCOM sur l'échantillonnage et l'analyse de l'eau de mer, comme indiqué dans les annexes I à III. Si l'une de ces lignes directrices convient dans le contexte de l'IMAP, elle peut être utilisée par les

[1] UNEP/MAP (2019). UNEP/MED WG.467/5. IMAP Guidance Factsheets: Update for Common Indicators 13, 14, 17, 18, 20 and 21: New proposal for candidate indicators 26 and 27.
[2] UNEP (2019a). UNEP/MED WG.463/6. Monitoring Protocols for IMAP Common Indicators related to pollution.

laboratoires méditerranéens compétents pour développer leurs propres méthodologies d'échantillonnage et de traitement des échantillons.

7. L'organigramme ci-dessous indique la catégorie de cette ligne directrice de surveillance relative à l'échantillonnage et à la conservation des échantillons d'eau de mer en vue de l'analyse de l'indicateur commun 17 de l'IMAP dans la structure de toutes les directives de surveillance élaborées pour les indicateurs communs 13, 14, 17, 18 et 20 de l'IMAP.



Organigramme : Lignes directrices pour la surveillance des objectifs écologiques 5 et 9 de l'IMAP

2 Note technique sur l'échantillonnage et le prétraitement de l'eau de mer en vue de l'analyse des métaux lourds[3] et des contaminants organiques

8. L'échantillonnage de l'eau de mer doit être effectué en même temps et au même endroit que l'échantillonnage d'autres matrices (sédiments, biotes) et que les mesures des effets biologiques (CIEM/OSPAR, 2012 [4]). L'échantillonnage, le prétraitement et l'analyse sont des activités complexes qui nécessitent une conception et une mise en œuvre minutieuses. En raison des très faibles concentrations en métaux lourds dans l'eau de mer (en particulier dans les stations de haute mer), une mauvaise manipulation de l'échantillon peut facilement entraîner une perte de déterminant et/ou une contamination de l'échantillon avant l'analyse. Des protocoles d'échantillonnage et de prétraitement appropriés constituent donc une étape cruciale de tout programme de surveillance de l'eau de mer.

9. La taille de l'échantillon d'eau de mer doit être suffisante pour prendre en charge les limites de détection souhaitées pour les contaminants concernés. Les lignes directrices du CIEM/OSPAR (2012) pour l'analyse de l'eau de mer (annexe I) suggèrent de recueillir un volume d'eau de mer approprié pour l'analyse en fonction de la concentration du contaminant dans la station spécifique (polluée ou non polluée) de telle sorte que la limite de quantification (LOQ) soit égale ou inférieure à une valeur de 30 % du critère d'évaluation pertinent (c'est-à-dire la norme de qualité environnementale, directive 2009/90/CE de la Commission [5]).

10. Il existe deux façons d'aborder l'analyse de l'eau de mer : a) à partir de l'eau de mer non filtrée et b) à partie de l'eau de mer filtrée. L'analyse d'échantillons d'eau non filtrée donne des résultats sur la concentration totale de contaminants dans l'eau de mer, quelles que soient les formes chimiques ou la taille des particules (c'est-à-dire dissoutes, complexées et liées à des colloïdes et à des particules en suspension (PS)). Par conséquent, des informations importantes sur la distribution et la disponibilité des contaminants sont perdues. D'autre part, la filtration par un tamis de 0,45 µm permet de séparer l'eau de mer filtrée (c'est-à-dire librement dissoute, complexée et liée), de la phase particulaire des contaminants, qui est retenue dans le filtre. Cependant, en raison des échanges de

[3] The term "heavy metals" is used indicating both heavy metals and trace elements
[4] ICES/OSPAR (2012). JAMP guideline on monitoring of contaminants in seawater: Annex 1: Guidelines for Monitoring of Contaminants in Seawater. ICES Advice 2012, Book 1
[5] EC (2009). Commission Directive 2009/90/EC laying down, pursuant to Directive 2000/60/EC of the European Parliament and of the Council, technical specifications for chemical analysis and monitoring of water status.

contaminants entre les formes chimiques des phases dissoutes et particulaires, ainsi que de l'influence potentielle des équipements d'échantillonnage et de filtration (filtres, parois des contenants, etc.), les équilibres entre les phases dissoutes et particulaires peuvent être modifiés au cours du processus. Par conséquent, la filtration doit être effectuée de manière à minimiser l'altération de l'échantillon d'eau de mer et la répartition des contaminants entre les phases dissoute et particulaire. De plus, dans le cas des contaminants organiques, leur répartition entre la phase dissoute et la phase particulaire est influencée par leur polarité, qui peut être exprimée par leur coefficient octanol/eau (log Kow ; Kow = concentration dans la phase octanol / concentration dans la phase aqueuse). Les composés les plus hydrophiles avec des valeurs de log Kow de 3 à 4 (tels que les aromatiques à 2 et 3 cycles et les isomères HCH) se trouvent principalement dans l'eau, tandis que les polluants avec des valeurs de log Kow > 5 (aromatiques à 4 à 6 cycles, groupe DDT, PCB) se trouvent dans les particules en suspension (PS). Des composés hydrophobes non polaires sont associés aux PS, qui sont séparées par filtration, mais ils sont également présents dans le filtrat adsorbé sur des colloïdes. En conséquence, la validation des procédures de séparation des phases est très difficile.

11. La filtration peut se faire en ligne (à partir de la bouteille d'échantillonnage ou du système de pompage de l'eau de mer) ou hors ligne, en laboratoire. Les systèmes de filtrage en ligne ont l'avantage de réduire le risque de perte de déterminant et/ou de contamination de l'échantillon provenant des bouteilles de stockage ou de l'air. Dans tous les cas, la filtration doit être autant que possible effectuée dans une zone exempte de particules. Il est préférable de travailler dans une hotte à flux laminaire. Les conditions recommandées pour un « banc stérile » ou un « laboratoire stérile » correspondent à la classe ISO 5 (GEOTRACES, 2017 [6]).

12. Des lignes directrices détaillées sur l'échantillonnage et le traitement de l'eau de mer figurent dans les documents publiés par le CIEM/OSPAR (2012) (I), HELCOM (2012a[7]) (annexe II), HELCOM (2012b[8]) (annexe III) et GEOTRACES (2017). En s'appuyant sur ces documents, sous cette note technique, les lignes directrices pour l'échantillonnage et la préservation des échantillons d'eau de mer pour l'indicateur commun 17 de l'IMAP fournissent les protocoles IMAP suivants pour l'échantillonnage de l'eau de mer :

- Protocole d'échantillonnage de l'eau de mer pour l'analyse des métaux lourds ;
- Protocole de filtration de l'eau de mer (métaux lourds) ;
- Protocole de stockage à bord d'échantillons d'eau de mer en vue de l'analyse des métaux lourds ;
- Protocole d'échantillonnage de l'eau de mer en vue de l'analyse des contaminants organiques ;
- Protocole de filtration de l'eau de mer (contaminants organiques) ;
- Protocole de stockage à bord d'échantillons d'eau de mer en vue de l'analyse des contaminants organiques.

2.1 Protocole d'échantillonnage de l'eau de mer en vue de l'analyse des métaux lourds

a) Matériel d'échantillonnage pour la collecte d'eau de mer

13. Habituellement, pour l'analyse des métaux, on recueille des échantillons d'eau de mer provenant de différentes profondeurs à l'aide de bouteilles GO-FLO (General Oceanics). L'échantillonneur se compose d'un cylindre muni d'un revêtement intérieur en téflon qui peut être abaissé en étant fermé dans la colonne d'eau et qui s'ouvre automatiquement par pression hydrostatique à une certaine profondeur. Cela permet d'éviter le contact de l'échantillon avec le film de surface de l'eau qui est enrichi en contaminants. D'autres types de bouteilles de prélèvement peuvent également être utilisés (comme les bouteilles Niskin), correctement modifiées pour éviter la contamination par les métaux.

[6] GEOTRACES (2017). Sampling and Sample-handling Protocols for GEOTRACES Cruises (Version 3), edited by the 2017 GEOTRACES Standards and Intercalibration Committee.
[7] HELCOM (2012a). Manual for marine monitoring in the COMBINE programme. Annex B-11, Appendix 1. Technical Note on the determination of trace metals (Cd, Pb, Cu, Co, Zn, Ni, Fe) including mercury in seawater.
[8] HELCOM (2012b). Manual for marine monitoring in the COMBINE programme. Annex B-11, Appendix 2. Technical note on the determination of persistent organic pollutants in seawater.

14. Tous les échantillonneurs doivent être nettoyés avant la première utilisation en rinçant les surfaces intérieures à l'aide d'acide chlorhydrique dilué. En pleine mer, les bouteilles doivent être rincées à l'eau de mer entre les prélèvements, tandis que dans les stations polluées, elles peuvent être rincées à l'eau déionisée.

15. La coque métallique du navire est une source potentielle de contamination en métaux (fer et plomb), tout comme l'utilisation de peintures anti-salissures (cuivre et étain) et la protection anodique du navire (zinc). Pour éviter la contamination par les métaux, le navire doit être positionné de manière à ce que la direction du vent et du courant marin minimisent l'influence de la coque du navire sur les échantillons d'eau de mer.

16. On utilise un équipement d'échantillonnage (tel qu'une bouteille GO-FLO ou de type Niskin d'une capacité de 12 à 30 l) fixé individuellement au câble hydrographique ou placé dans un système de rosette sans métal (figure 1).

Figure 1. Échantillonneur d'eau de mer individuel GO FLO et système de rosette avec plusieurs échantillonneurs

17. Le câble hydrographique doit être en acier inoxydable recouvert de téflon, en polymère ou en Kevlar pour éviter la contamination par les métaux. Tous les poids servant de lest pour abaisser les bouteilles/rosettes doivent être non métalliques ou recouverts de résines époxy pour éviter la contamination par les métaux

i) Les bouteilles de prélèvement sont descendues aux profondeurs définies. Un enregistreur de profondeur est installé sur les bouteilles individuelles ou sur le système de rosette pour contrôler la profondeur d'échantillonnage ;

ii) On utilise un messager non métallique (ou recouvert de résines époxy) pour libérer les vannes de fermeture aux deux extrémités de l'échantillonneur pour les bouteilles individuelles ou un système de déclenchement afin de fermer les bouteilles dans un système de rosette lors de l'ascension ;

iii) Une fois que les échantillons d'eau de mer ont été prélevés à toutes les profondeurs, le système de bouteilles/rosettes est hissé à bord ;

iv) L'équipement d'échantillonnage doit alors être placé dans un sac en plastique ou un autre contenant pré-nettoyé, puis transporté dans une zone ISO de classe 5 (ou une hotte avec de l'air filtré sans métal) pour une manipulation ultérieure ;

v) Les échantillons d'eau de mer sont transférés des bouteilles GO-FLO (ou d'un équipement d'échantillonnage similaire) dans des bouteilles en téflon (ou en polyéthylène) pré-nettoyées (avec du HCl ou du HNO3 dilué) pour l'analyse des métaux totaux ;

vi) Dans le cas où les PS sont analysés séparément de la fraction métallique dissoute, l'échantillon d'eau de mer est transféré à l'unité de filtration, à l'aide d'un tube en téflon pré-nettoyé.

18. La contamination des échantillons par l'atmosphère (comme la peinture et les particules de rouille, les gaz d'échappement des moteurs et le fond atmosphérique) peut être très importante et des mesures doivent être prises pour l'éviter. Par conséquent, la manipulation de l'eau de mer doit se faire dans un environnement sans poussière et sans métal, dans des conditions contrôlées (zone ISO de classe 5).

19. On utilise notamment des gants en latex ou en nitrile neutres pour manipuler les échantillons d'eau de mer afin d'éviter toute contamination.

b) Pompage d'eau de mer in situ (profils)

20. Le pompage *in situ* de l'eau de mer à des profondeurs données est une méthode alternative de collecte de l'eau de mer, qui minimise la manipulation de l'échantillon susceptible d'entraîner la perte des déterminants et/ou la contamination de l'échantillon par l'air. Le système de pompage peut éventuellement inclure une filtration en ligne, pour séparer les PS du filtrat d'eau de mer. La méthode peut être utilisée pour des profondeurs relativement faibles (jusqu'à 100 m) en utilisant une pompe péristaltique ou un piston en téflon, ou des pompes à membrane et des tubes en silicone, polyéthylène ou téflon, afin d'éviter la contamination par les métaux. Avant l'utilisation, la tubulure doit être nettoyée en pompant de l'acide dilué (tel que HCl ou HNO_3). Lors de l'échantillonnage, les premiers litres d'eau de mer doivent être rejetés afin de rincer tout le système de pompage avant la collecte des échantillons d'eau de mer. Le volume de rinçage dépend de la longueur du tuyau utilisé et il faut rincer en utilisant au moins trois fois le volume du tuyau avant de prélever l'échantillon proprement dit. Avant son utilisation sur le terrain, le fonctionnement et les performances de la pompe doivent être soigneusement vérifiés et optimisés. (Figure 2).

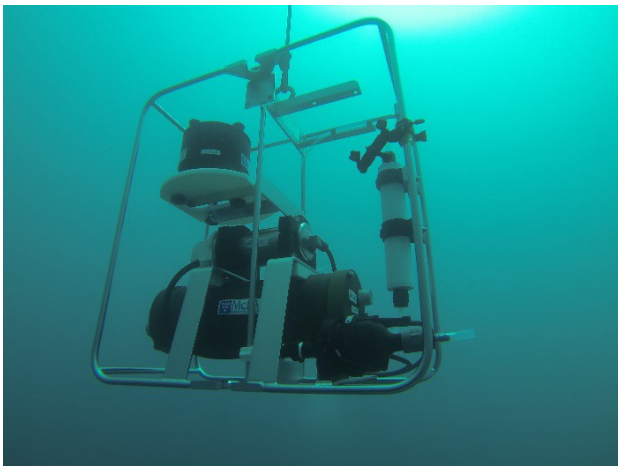


Figure 2. Système de pompage d'eau de mer *in-situ*
(Laboratoire d'études environnementales marines, AIEA)

21. Le flux sortant du système de pompage est collecté dans des bouteilles sans métal (polyéthylène, téflon, verre). Pour l'analyse du mercure, l'eau doit être recueillie dans des bouteilles en verre ou en quartz. Si un système en ligne est attaché au dispositif de pompage, le filtrat doit être stocké dans des bouteilles sans métal (comme ci-dessus), tandis que les filtres avec les échantillons de PS doivent également être placés dans des récipients sans métal.

c) Échantillonnage *in situ* de l'eau de mer de surface

22. Pour l'échantillonnage de l'eau de mer de surface, GEOTRACES (2017) recommande un système de pompage de surface de type siphon/poisson remorqué qui consiste en :

i) Une pompe à diaphragme en téflon-PTFE avec tube de pompe en silicone ;

ii) Des tubes d'échantillonnage en téflon-PFA ;

iii) Une ailette de dépression en PVC de 1 m au-dessus d'un poids de 20 kg, logé dans un poisson en PVC (ou un poisson en acier inoxydable de 50 kg) qui ne nécessite pas de dépresseur séparé ;

iv) Une ligne tressée en polyester reliant le poisson au dépresseur (si nécessaire) puis au navire ; le tube d'échantillonnage en téflon est déplacé le long de cette ligne ;

v) Un tube en téflon PFA est utilisé de l'autre côté de la pompe pour acheminer l'eau de mer directement dans une zone propre pour l'échantillonnage ;

vi) Pour un échantillonnage d'eau de surface en à des vitesses de 1 à 12 nœuds, le système de

pompage *sipper* est déployé sur le côté du navire en utilisant la grue du navire pour suspendre le poisson à l'extérieur du sillage de la proue, la prise d'eau se trouvant à environ 2 m de profondeur. Ce modèle de siphon permet d'atteindre des vitesses plus rapides s'il y a peu ou pas de houle et si le siphon reste à l'écart des vagues de proue. La conception du siphon permet également un échantillonnage quasi-stationnaire (avançant dans l'eau propre à une vitesse de 0,5 à 1 nœud) afin de recueillir de grands volumes d'eau de mer sans métaux-traces à des profondeurs pouvant atteindre 25 m.

Figure 3. Système de pompage de surface / de poisson remorqué (GEOTRACES, 2017)

d) Nettoyage de l'équipement et du matériel de laboratoire avant le prélèvement des échantillons

23. Un protocole de nettoyage de l'équipement de prélèvement d'échantillons d'eau de mer pour l'analyse des métaux est proposé par HELCOM (2012a) (Annexe II)

i) Les instruments de laboratoire sont stockés dans du HCl 2M (haute pureté) pendant une semaine, rincés à l'eau, conservés dans l'eau pendant une semaine et séchés dans des conditions exemptes de poussière (banc stérile).

ii) Les dispositifs d'échantillonnage sont remplis de HNO_3 (haute pureté) à 1 %, conservés à température ambiante pendant trois semaines et rincés à l'eau.

iii) Les bouteilles en téflon/quartz sont stockées dans du HCl chaud (40°C, ±5°C) dilué au 1 :1 pendant une semaine. Elles sont ensuite rincées à l'eau et stockées avec du HNO_3 1M (haute pureté) jusqu'à leur utilisation finale (trois semaines minimum).

24. Des procédures de nettoyage modifiées sont nécessaires pour le mercure. Les récipients en verre (borosilicate, quartz) utilisés pour la collecte et le stockage des échantillons en vue de la détermination de la concentration de mercure sont généralement nettoyés en utilisant une procédure d'oxydation décrite par Sturgeon et Berman (1987 [9]). Les bouteilles sont remplies d'une solution de 0,1 % de $KMnO_4$, 0,1 % de $K_2S_2O_8$ et 2,5 % de HNO_3, et chauffées pendant 2 heures à 80 °C. Les bouteilles sont ensuite rincées à l'eau et stockées avec 2 % de HNO_3 contenant 0,01 % de $_2Cr_2O_7$ ou de $KmnO_4$ jusqu'à ce qu'elles soient prêtes à être utilisées.

25. Des protocoles détaillés de nettoyage du matériel d'échantillonnage et des bouteilles de stockage sont également proposés par GEOTRACES (2017) et CIEM/OSPAR (2012) (annexe I).

2.2 Protocole de filtration de l'eau de mer en vue de l'analyse des métaux lourds

a) Procédure de filtrage : L'eau de mer doit être filtrée dès que possible après le prélèvement des échantillons, pour éviter toute variation du rapport entre la concentration de contaminants dissous et de contaminants sous forme de particules.

Filtration en ligne

26. L'eau de mer peut être directement filtrée à partir de bouteilles GO-FLO pressurisées en utilisant une faible surpression (<50 kPA, ou <7 psi, maximum) d'azote gazeux filtré de haute qualité ou d'air comprimé pour obtenir un débit suffisant à travers les filtres (GEOTRACES, 2017). Avant de commencer la filtration, il est recommandé d'agiter légèrement les bouteilles GO-FLO car les

[9] Sturgeon, R., and Berman, S. 1987. Sampling and storage of natural water for trace metals. In Critical reviews in Analytical Chemistry. 18(3): 209-244. CRC Press.

particules peuvent se déposer entre la fermeture de la GO-FLO en profondeur et le début de la filtration. Un filtre à capsule ou un porte-filtre à membrane pré-nettoyé est connecté à la vanne à clapet en téflon de la GO-FLO avec un tube en téflon-PFA (ou équivalent propre) et les bouteilles d'échantillon sont remplies avec l'effluent de ce filtre (les filtres à capsule doivent être rincés avec environ 0,5 L d'eau d'échantillon avant la collecte du filtrat).

Filtration hors ligne

27. Après avoir été prélevée dans les bouteilles de prélèvement GO-FLO, l'eau de mer est transférée dans une bouteille secondaire, d'où elle est envoyée vers l'équipement de filtration. La filtration hors ligne donne des résultats similaires à la filtration en ligne si des procédures strictes de travail sans traces de métaux sont respectées. Elle peut donc être utilisée si les limites de manipulation des échantillons à bord du navire de prélèvement l'exigent. Avant de commencer la filtration, il est recommandé d'agiter légèrement les bouteilles GO-FLO car les particules peuvent se déposer entre la fermeture de la GO-FLO en profondeur et le début de la filtration. Ensuite, on draine l'eau de mer dans une bouteille de transfert pré-nettoyée, qui est bouchée et transférée dans la zone de filtration. Le volume à filtrer doit être de 5-10 l, ce qui est suffisant pour charger les filtres avec suffisamment de matière pour dépasser les blancs de filtre pour presque tous les échantillons et tous les analyses (GEOTRACES, 2017).

28. Une fois la filtration terminée, l'eau de mer résiduelle peut être forcée à travers le filtre à l'aide d'une seringue en polypropylène remplie d'air. Cela permet d'éviter le déversement et la perte de particules depuis la face du filtre lorsque le porte-filtre est ouvert. Les porte-filtres peuvent ensuite être démontés et les filtres retirés avec précaution à l'aide de pinces en téflon et stockés dans des lamelles de Pétri ou dans un récipient approprié similaire, puis congelés à -20° C.

b) Filtres

29. Les filtres en polycarbonate (0,45 µm) sont couramment utilisés pour la filtration de l'eau de mer en vue de l'analyse des métaux lourds (sauf le mercure). Le choix d'un filtre se fait en fonction de plusieurs critères faible teneur en métal, résistance mécanique, facilité de manipulation, capacité de charge en particules relativement élevée, faible tendance à se colmater complètement et qualité du débit de filtration. Les filtres doivent être nettoyés avec du HCl 2M (haute pureté) pendant au moins trois semaines, rincés à l'eau déionisée et stockés dans l'eau pendant une semaine supplémentaire (HELCOM, 2012a). Ils doivent ensuite être séchés sur un banc stérile et stockés dans un dessiccateur jusqu'à ce que leur poids soit constant. La même procédure de séchage et de pesage doit être appliquée aux filtres chargés de PS (Pohl, 1997 [10]).

30. Pour le dosage du mercure, il est recommandé d'utiliser des filtres en fibre de verre (qualité GF/F, type Millipore) et des filtres en téflon. La procédure de nettoyage de ces filtres est comparable à celle utilisée pour les filtres en polycarbonate (Queremais et Cossa 1997 [11]).

31. Le diamètre des filtres dépend de la quantité de particules en suspension dans les stations d'échantillonnage. Alors qu'un diamètre de filtre de 25 mm est suffisant pour filtrer 10 l d'eau de mer sans colmatage dans les stations en pleine mer, un format de 47 mm est préférable pour les stations du plateau/talus continental où les concentrations de particules sont plus élevées. On s'efforce de minimiser le diamètre du filtre afin de maximiser la charge en particules par surface de filtre, et donc de réduire le blanc du filtre par rapport aux concentrations en métaux dans les PS.

[10] Pohl, C. 1997. Trace Metals (Cu, Pb, Zn, Cd, Al, Li, Fe, Mn, Ni, Co) in Marine Suspended Particulate Matter: An International ICES Intercomparison Exercise. Accreditation and Quality Assurance, 2: 2-10

[11] Quémerais, B., and Cossa, D. 1997. Procedures for sampling and analysis of mercury in natural waters. Environment Canada-Quebec region, Environmental Conservation, St. Lawrence Centre. Scientific and Technical Report ST-31E, 34 pp.

32. Les porte-filtres en polypropylène sont couramment utilisés car ils sont compatibles avec les procédures de travail propres sans traces de métaux. Il est important d'avoir des capacités d'étanchéité parfaites sous pression.

c) Nettoyage des filtres et des porte-filtres

33. GEOTRACES (2017) propose le protocole suivant pour le nettoyage des filtres et porte-filtres pour l'analyse des métaux traces dans les échantillons d'eau de mer :

i) Une bouteille pré-nettoyée de 1000 ml en polyéthylène basse densité (PEBD) peut être pré-nettoyée à nouveau en la remplissant de 10 % (v/v, ou 1,2 M) de HCl de qualité TM, en la mettant en double sac dans des sacs en polyéthylène Ziploc résistants (par exemple de 4 mm) et en la plaçant dans un four à 60 °C pour une durée allant de 4 heures à une nuit entière.
ii) La bouteille est retirée de la hotte et placée à l'envers de sorte que le couvercle soit lessivé par l'acide pendant que celui-ci refroidit. L'acide est évacué et la bouteille est rincée abondamment au moins 3 fois avec de l'eau déionisée sans métaux-traces (par exemple, Milli- Q).
iii) La bouteille propre est remplie à 90 % d'eau déionisée sans métaux-traces.

34. Les filtres doivent être retirés de la boîte d'origine à l'aide d'une pince exempte de métal, permettant de les saisir par les bords afin de ne pas endommager la région de l'échantillon, et déposés avec précaution dans le flacon. Il faut veiller à ce que les feuilles de papier séparant les filtres dans l'emballage d'origine ne soient pas inclus. Lorsque 100 filtres ont été immergés dans l'eau, les derniers 10 % du volume de la bouteille sont remplis de HCl concentré de qualité sans métaux-traces. La bouteille est fermée hermétiquement, mélangée délicatement pour que les filtres ne se froissent pas, puis elle est mise dans un double sac et placée dans un four à 60°C pendant la nuit, comme pour le nettoyage des bouteilles.

35. Lorsque la bouteille de filtres est froide, l'acide est lentement évacué, tout en retenant les filtres avec le bouchon maintenu contre le goulot de la bouteille. Les filtres sont maintenus en suspension par une légère agitation manuelle pendant que l'acide est évacué, afin de minimiser le pliage et le froissement pendant que toute la solution est éliminée. On remplit lentement la bouteille d'eau déionisée que l'on fait couler lentement le long de la paroi intérieure, tout en agitant doucement, puis l'eau est évacuée, tout en retenant les filtres avec le bouchon. La procédure est répétée 5 fois. On garde le dernier rinçage dans la bouteille et on laisse reposer à température ambiante toute la nuit afin que tout acide résiduel se diffuse par les pores des filtres. Trois autres rinçages sont répétés le lendemain. Il faut toujours vérifier le pH pour s'assurer qu'il ne reste pas d'acide, car les filtres peuvent nécessiter de nombreux rinçages avant élimination de toute trace d'acide. Les filtres peuvent être laissés en suspension dans l'eau déionisée jusqu'à leur utilisation sur le navire, ou peuvent être chargés à l'avance dans des lamelles de Pétri individuelles pour un accès et un stockage facile dans la même lamelle à Pétri. Il faut être prudent pour éviter d'avoir des filtres doublés, car les filtres ont tendance à se coller les uns aux autres (GEOTRACES, 2017).

2.3 Protocole de stockage à bord d'échantillons d'eau de mer en vue de l'analyse des métaux lourds

36. Les échantillons d'eau de mer doivent être stockés dans des conditions permettant d'éviter toute perte ou contamination de métaux pendant le transfert depuis le navire vers le laboratoire, pour prétraitement et analyse. Le processus couramment utilisé pour la conservation des échantillons d'eau de mer en vue de l'analyse des oligo-éléments est l'acidification et la congélation. Toutefois, tout sous-échantillonnage des échantillons d'eau de mer doit être effectué à bord, immédiatement après l'échantillonnage. Si une filtration est nécessaire pour séparer les PS de la phase dissoute de l'échantillon, elle doit également être effectuée immédiatement après l'échantillonnage et avant toute addition d'acide pour des raisons de conservation.

37. Les échantillons d'eau de mer (filtrée ou non filtrée, si les métaux totaux doivent être analysés) sont acidifiés par l'ajout de 1,5 ml de HNO_3 ou de HCl (haute pureté) par litre d'échantillon d'eau de mer immédiatement après la filtration, pour une acidification à un pH de 1,0 à 1,6. Les bouteilles sont stockées à 4ºC dans l'obscurité. Les filtres avec des PS doivent être conservés dans des boîtes en plastique à -20 °C. Dans ces conditions, les échantillons d'eau et les matières particulaires sur les filtres peuvent être conservés pendant au moins un an. Pour l'analyse du Hg, il faut ajouter des

agents d'oxydation d'acidification (tels que le $Cr_2O_7^{2-}$). (HELCOM, 2012a)

38. Les bouteilles utilisées pour le stockage de l'eau de mer doivent être en polyéthylène basse densité (PEBD) ou en polyéthylène haute densité (HPDE). Les bouchons des bouteilles sont généralement en polypropylène, un matériau adapté au stockage de l'eau de mer. Pour le Hg, les bouteilles en polyéthylène ne sont pas recommandées et on peut utiliser à la place des bouteilles en verre ou en téflon (GEOTRACES, 2017).

a) Nettoyage de bouteille d'échantillonnage

39. Le nettoyage des bouteilles utilisées pour le stockage des échantillons d'eau de mer en vue de l'analyse des éléments traces, doit être très minutieux pour éviter l'altération de l'échantillon à partir du récipient. GEOTRACES (2017) propose un protocole très rigoureux pour le nettoyage des bouteilles ; il est utilisé par des groupes de recherche ayant une longue expérience de réussite de l'échantillonnage propre des métaux traces. Les protocoles GEOTRACES sont les suivants :

Protocole GEOTRACES pour les bouteilles en PEBD et PEHD (éléments traces dissous et solvables) :

i) Les bouteilles doivent parfois être rincées avec du méthanol ou de l'acétone pour libérer les huiles de fabrication.

ii) Faire tremper les bouteilles pendant une semaine dans un détergent alcalin (par exemple Micro, Decon). Ce processus peut être accéléré par un trempage à 60°C pendant une journée

iii) Rincer 4 fois par osmose inversée ou avec de l'eau déminéralisée.

iv) Rincer 3 fois avec de l'eau ultra-pure dans une atmosphère propre.

v) Remplir les bouteilles avec du HCl 6M (qualité réactif) et les immerger dans un bain de HCl 2M (qualité réactif) pendant un mois. Là encore, on peut accélérer le processus en chauffant pendant une semaine.

vi) Rincer 4 fois avec de l'eau ultra-pure dans une atmosphère propre.

vii) Remplir les bouteilles avec du HCl 1 M (qualité métal-trace) pendant au moins un mois. Stockage dans un double sac. Les bouteilles doivent être vidées de tout acide avant leur transport vers le navire.

viii) Rincer avec de l'eau ultra-pure et embarquer les bouteilles vides dans des double sacs.

Protocole GEOTRACES pour les bouteilles en téflon PFA :

i) Faire tremper les bouteilles pendant une journée dans un détergent alcalin ;
ii) Rincer 7 fois avec de l'eau déionisée jusqu'à ce qu'il n'y ait plus de trace de détergent ;
iii) Rincer 3 fois avec de l'eau ultra-pure ;
iv) Faire tremper dans un bain de HCl de qualité réactif 6 M pendant 1 journée ;
v) Rincer 5 fois avec de l'eau ultra-pure ;
vi) Remplir les bouteilles d'acide nitrique 1M (qualité analytique) et les conserver à 100°C pendant 5 heures sous une hotte ;
vii) Rincer 5 fois avec de l'eau ultra-pure à l'intérieur d'une hotte à flux laminaire de classe ISO 5 ;
viii) Remplir les bouteilles avec de l'eau ultra-pure et les conserver à 80°C pendant 5 heures ;
ix) Rincer 5 fois avec de l'eau ultra-pure à l'intérieur d'une hotte à flux laminaire de classe ISO 5. Stockage dans un double sac.

2.4 Protocole d'échantillonnage de l'eau de mer en vue de l'analyse des contaminants organiques

a) Matériel d'échantillonnage pour la collecte d'eau de mer

40. Les concentrations de contaminants organiques dans l'eau de mer sont généralement très faibles, c'est pourquoi, pour atteindre la limite de quantification (LQ) requise pour ces contaminants (en pg l^{-1}), il faut collecter de grands volumes d'eau (parfois plus de 100 litres) à extraire pour éviter les interférences du fond de la matrice (CIEM/OSPAR, 2012) (annexe I). Cependant, de grands volumes d'eau de mer ne peuvent pas être facilement manipulés et transportés, c'est pourquoi

l'extraction d'eau de mer à bord résout beaucoup de problèmes de logistique et évite l'altération des caractéristiques des échantillons d'eau de mer. L'équipement de filtration/extraction *in situ* présente en outre l'avantage d'une courte exposition de l'échantillon d'eau de mer à l'atmosphère.

41. Pour l'échantillonnage de l'eau de mer ou l'analyse des contaminants organiques, les équipements sont de préférence en verre ou en acier inoxydable. Des équipements recouverts de téflon peuvent également être utilisés pour les composés organiques persistants et les HAP.

42. Les bouteilles en verre sont un équipement d'échantillonnage approprié pour l'analyse des contaminants organiques. Les bouteilles sont montées dans une cage en acier inoxydable et sont descendues sur un câble hydrographique jusqu'à la profondeur d'échantillonnage souhaitée, ouvertes sous l'eau puis remontées sur le pont du navire. L'échantillonneur en verre peut être utilisé jusqu'à une profondeur de 2000 m (10 l) et 100 m (100 l) (CIEM, 2012) (figure 4). Pour une profondeur plus importante, il est possible d'utiliser des bouteilles en acier inoxydable, basées sur le design Niskin et GO-FLO. Un système d'enregistrement de la profondeur est installé sur le boîtier en acier, pour permettre la collecte de l'eau de mer à la profondeur souhaitée.

Figure 4. Bouteille en verre pour l'échantillonnage de l'eau de mer en vue de l'analyse des contaminants organiques (ICES/OSPAR 2012)

43. Tous les échantillonneurs doivent être nettoyés avant la première utilisation, à l'aide des solvants organiques appropriés. En pleine mer, les bouteilles doivent être rincées à l'eau de mer entre les prélèvements, tandis que dans les stations polluées, elles peuvent être rincées à l'eau déionisée.

44. Une fois que l'équipement d'échantillonnage est hissé à bord, il doit être placé immédiatement dans un conteneur en aluminium ou en acier inoxydable et transporté dans une salle blanche (ou une hotte avec de l'air filtré exempt de poussière) dans le laboratoire du navire, pour une manipulation ultérieure. La contamination des échantillons par l'atmosphère (comme les HAP provenant des gaz d'échappement des moteurs) ou par le navire (c'est-à-dire les PCB dans l'huile de lubrification) peut entraîner une contamination des échantillons, c'est pourquoi des mesures doivent être prises pour l'éviter, comme par exemple le positionnement du navire par rapport à la direction du vent et du courant marin afin de minimiser tout risque de contamination liée au navire. Toute manipulation d'eau de mer doit être effectuée dans un environnement sans poussière, dans des conditions contrôlées.

b) Échantillonnage par pompage - filtration et extraction *in situ*

45. Le pompage *in situ* de l'eau de mer à des profondeurs déterminées constitue une méthode alternative de collecte de l'eau de mer, qui minimise la manipulation de l'échantillon susceptible d'entraîner la perte de déterminants et/ou la contamination de l'échantillon par l'air. Le système de pompage peut éventuellement inclure une filtration en ligne, pour séparer les PS du filtrat d'eau de mer. La filtration *in-situ* suivie d'une extraction en phase solide minimise le risque de contamination de l'échantillon pendant le prélèvement. Le système de pompage comprend un filtre en fibre de verre (taille des pores 0,7 µm) pour recueillir la phase particulaire et une colonne de verre garnie de résine polymère pour la phase dissoute. Le système de pompage fonctionne de la même manière que pour l'analyse des métaux lourds (paragraphe 17). Des volumes de 1 à 100 l peuvent être échantillonnés par

échantillonnage et/ou pompage discret et sont généralement extraits soit par extraction liquide-liquide (ELL) soit par extraction en phase solide (EPS), tandis que les volumes plus importants sont généralement échantillonnés par pompage et extraits par extraction en phase solide (CIEM/OSPAR, 2012).

46. Des informations détaillées sur l'étalonnage du système de pompage *in situ* sont fournis par le fabricant. Avant son utilisation sur le terrain, le fonctionnement et les performances de la pompe doivent être soigneusement vérifiés et optimisés.

2.5 Protocole de filtration de l'eau de mer en vue de l'analyse des contaminants organiques

a) Procédure de filtrage/extraction

47. Les concentrations de contaminants organiques dans l'eau de mer sont très faibles (les LOQ sont de l'ordre de pg l^{-1}). C'est pourquoi de grands volumes d'eau (10 à 100 l ou plus) doivent être filtrés et extraits pour surmonter les problèmes de blancs. Comme les composés hydrophobes se présentent sous des formes dissoutes, colloïdales et liées aux particules, la filtration doit être effectuée de manière à éviter l'altération de la répartition des composés organiques entre les phases dissoute et particulaire en raison de la manipulation des instruments. Il est donc préférable que la filtration soit effectuée immédiatement après le prélèvement.

Filtration/extraction in situ

48. Afin de minimiser l'altération de la répartition des contaminants organiques entre les phases, ainsi que la contamination de l'air, la filtration/extraction *in situ* peut être effectuée à l'aide d'une pompe à eau submersible. La filtration/extraction in-situ est compacte et présente le double avantage d'un équipement de petite taille et d'une exposition brève à l'atmosphère (HELCOM, 2012b). La pompe, qui comprend un porte-filtre, une colonne de résine polymère, une pompe et un débitmètre, est déployée à une profondeur déterminée sur un câble hydrographique et le pompage est lancé et arrêté par télécommande. Un filtre en fibre de verre (taille des pores 0,7 µm) récupère la phase particulaire et une colonne en verre garnie de résine polymère récupère la phase dissoute. Étant donné que les pompes submersibles comportent généralement des pièces et des raccords en plastique, il convient de vérifier avant utilisation si la pompe ne contient pas de contaminants organiques ciblés, afin de remplacer si nécessaire les pièces concernées par des pièces en acier inoxydable ou en verre (si possible) afin de réduire la contamination. Des étalons de substitution peuvent être ajoutés à la colonne de résine avant l'échantillonnage pour contrôler les récupérations de l'extraction et le stockage. La méthode d'échantillonnage par pompe *in-situ* doit être validée avant son utilisation (CIEM/OSPAR, 2012).

Filtration hors ligne

49. Le stockage des échantillons d'eau de mer pour la détermination des contaminants organiques n'est pas pratique en raison des grands volumes d'eau de mer nécessaires à la quantification des déterminants. En outre, la période de stockage des échantillons d'eau de mer avant extraction doit être limitée (moins de 2 heures, HELCOM, 2012b) et il est recommandé d'extraire l'échantillon d'eau le plus rapidement possible après l'échantillonnage. De plus, il est préférable d'éviter le transfert d'eau de mer dans un autre récipient, ainsi que les manipulations inutiles pouvant entraîner une altération des caractéristiques de l'échantillon. Les bouteilles de prélèvement doivent être soigneusement déplacées vers la zone propre du laboratoire de bord (Protocole IMAP 2.4. sur l'échantillonnage de l'eau de mer en vue de l'analyse des contaminants organiques) pour procéder à la filtration et à l'extraction.

50. Les bouteilles de prélèvement sont reliées à un filtre en fibre de verre (taille des pores 0,7 µm) pour récupérer la phase particulaire et la phase dissoute est extraite à bord par extraction liquide-liquide (ELL) ou extraction en phase solide (EPS). Les extraits ou les cartouches absorbantes sont stockés dans des conditions fraîches (< 4°C) et à l'abri de la lumière.

b) Filtres

51. La filtration se fait à l'aide de filtres en fibre de verre (GF/F) (taille de pore de 0,7 µm). Les filtres à lit plat ont une capacité très limitée, c'est pourquoi les filtres en fibre de verre enroulés sont souvent utilisés pour les volumes supérieurs à 10 l et les échantillons d'eau contenant de grandes

quantités de matières en suspension. Une pompe est nécessaire pour faire passer l'eau à travers le filtre (HELCOM, 2012b).

c) Nettoyage des filtres et des porte-filtres

52. Dans de nombreux cas, la limite de détection de la procédure est déterminée par la valeur du blanc. Afin de maintenir la valeur du blanc aussi basse que possible, les composés à analyser ou d'autres composés interférents doivent être éliminés des filtres et de tous les instruments en verre et les tubes utilisés pour la filtration.

53. Une procédure de nettoyage de tous les équipements et matériaux utilisés dans la manipulation et le traitement des échantillons d'eau de mer pour l'analyse des contaminants organiques est proposée par HELCOM (2012b) :

i) Les instruments en verre doivent être soigneusement lavés à l'aide des détergents et rincés au moyen d'un solvant organique avant d'être utilisés. Un nettoyage supplémentaire des objets en verre, autre que les instruments calibrés, peut être effectué par chauffage à des températures > 250 °C.
ii) Tous les solvants doivent être contrôlés pour détecter les impuretés en concentrant la quantité normalement utilisée à 10 % du volume final normal. Ce concentré est ensuite analysé de la même manière qu'un échantillon par HPLC ou GC et ne doit pas contenir de quantités significatives des composés à analyser ou d'autres composés interférents.
iii) Tous les produits chimiques et matériaux d'adsorption doivent être vérifiés pour détecter les impuretés et, le cas échéant, être purifiés (par exemple par chauffage ou extraction). Les cartouches Soxhlet doivent être pré-extraites. Les cartouches d'extraction en fibre de verre sont préférables à celles en cellulose. On peut aussi utiliser des cartouches Soxhlet en verre plein, avec un filtre en verre à gros rendement G1 disposé au fond. Le stockage de ces matériaux ultra propres pendant une longue période n'est pas recommandé, car l'air des laboratoires peut contenir des HAP qui seront adsorbés par ces matériaux. Les valeurs de blanc qui se produisent malgré toutes les précautions susmentionnées peuvent être dues à la contamination de l'air.

54. Comme les concentrations en HAP et en hydrocarbures chlorés dans l'eau de mer sont très faibles, il est très difficile de contrôler les problèmes de blancs et de contamination. Il est donc recommandé, le cas échéant, de laver à nouveau tout le matériel (flacons, pipettes, bouteilles en verre) avec du solvant juste avant l'utilisation. Les étapes critiques doivent être réalisées sur un banc stérile si possible.

2.6 Protocole de stockage à bord d'échantillons d'eau de mer en vue de l'analyse des contaminants organiques

55. L'eau de mer peut être stockée dans des bouteilles en verre afin d'éviter la contamination et de minimiser l'adsorption des contaminants organiques à la surface de la bouteille. Cependant, comme les composés très lipophiles tels que les HAP à 4 ou 6 cycles, le DDT, les PCB, ont tendance à adsorber sur toutes les surfaces, les échantillons doivent être extraits le plus rapidement possible après le prélèvement. La meilleure procédure consiste à extraire les échantillons par extraction liquide- liquide (ELL) ou par extraction en phase solide (EPS) et à stocker les extraits ou les cartouches d'adsorbant dans des conditions froides (< 4°C) et à l'abri de la lumière. Les extraits dans les solvants organiques sont moins susceptibles d'être adsorbés sur les surfaces (HELCOM, 2012b). Dans le cas ou des échantillons d'eau de mer doivent être stockés, la conservation doit également se faire à l'abri de la lumière et dans un réfrigérateur (à 4°C) (CIEM/OSPAR, 2012).

56. Les échantillons de matières particulaires en suspension après filtration doivent être réfrigérés (-20 °C) et conservés congelés pour être analysés ultérieurement.

Annex I

ICES/OSPAR (2012). JAMP guideline on monitoring of contaminants in seawater: Annex 1: Guidelines for Monitoring of Contaminants in Seawater. ICES Advice 2012, Book 1, (3.1.1)

1.5.5.4 Special request, Advice May 2012

ECOREGION General advice
SUBJECT Development of a JAMP guideline on monitoring of contaminants in seawater

Advice summary

ICES has developed a guideline document on monitoring of contaminants in seawater under the Joint Assessment and Monitoring Programme (JAMP) (Annex 1). The document also includes a technical annex on specifics of suitable sampling equipment. ICES advises that the document is included in the JAMP guidelines.

Request

Development of a JAMP guideline on monitoring of contaminants in seawater (OSPAR 2011/1)

To develop the general text for a JAMP guideline on monitoring contaminants in seawater, which could act as the overarching chapeau to technical annexes concerning specific substances. The technical annex on analysis of PFC compounds in seawater developed by ICES in 2009 is the first such document. The development of the overarching text should take into account the need to address the following issues: purposes; quantitative objectives; sampling strategy; sampling equipment; storage and pre-treatment of samples; analytical procedures; analytical quality assurance; reporting requirements.

ICES advice

ICES has developed guidelines for monitoring of contaminants in seawater (Annex 1), complementing the corresponding JAMP Guideline for Monitoring of Contaminants in Sediment and JAMP Guideline for Monitoring of Contaminants in Biota. The guideline document in Annex 1 covers monitoring for organic contaminants and trace metals and is structured along the sections outlined in the request (purposes, quantitative objectives, sampling strategy, sampling equipment, storage and pre-treatment of samples, analytical procedures, analytical quality assurance, and reporting requirements). In addition, an annex to the guideline has been developed on technical specifics of the sampling equipment suitable for subsequent analysis of organic contaminants and trace metals. The document includes references to the EU Water Framework Directive (WFD) and EU Marine Strategy Framework Directive (MSFD) where applicable.

ICES advises that this document is included in the JAMP guidelines.

Source

ICES. 2012. Report of the Marine Chemistry Working Group (MCWG), 20–24 February 2012, Southampton, UK. ICES CM 2012/SGHIE:05.

Annex 1: Guidelines for Monitoring of Contaminants in Seawater

1. Introduction

These guidelines provide advice on the sampling and analysis of seawater, for determination of trace metals and organic contaminants, including oceanic, coastal, and estuarine waters. Monitoring contaminants in seawater is a complex task which requires carefully designed and conducted sampling campaigns, appropriate sampling equipment and its correct handling, as well as suitable pre-treatment and storage methods for the analytes in question. There are numerous steps that will affect data quality prior to the chemical analysis itself.

Contaminants in seawater can originate from direct point sources, riverine discharges, and atmospheric dry and wet deposition. Their distribution in seawater depends on the physical-chemical characteristics of the compound or element, interactions with the water matrix, sediment and biota as well as hydrographical conditions, such as mixing of water masses. Organic contaminants and metals can occur freely dissolved in water, bound to colloids, or suspended particulate matter. Trace metals can form complexes with organic or inorganic material. This partitioning is the result of environmental conditions and the partitioning may change during sampling and storage, and has implications for analysis and interpretation.

These guidelines are general recommendations on contaminant monitoring in seawater. The techniques described are useful for routine monitoring and ship/campaign-based work. However, this guideline is not intended as a complete laboratory manual. Requirements for specific contaminants or contaminant groups should be further specified by expert groups, for example in associated technical annexes, in order to meet the objectives of the monitoring programme and to ensure consistent and comparable data sets.

2. Purposes

Monitoring of contaminants in seawater of the Northeast Atlantic Ocean is performed within the framework of OSPAR as the regional convention for the protection of the marine environment of this area. OSPAR monitoring also can assist member states of the European Union to fulfil their obligations under the relevant EU directives, such as the Marine Strategy Framework Directive (MSFD) (EU, 2008) and the Water Framework Directive (WFD) (EU, 2000) with its related directives such as the daughter directive on Environmental Quality Standards in the field of water policy (2008/105/EC).

One of the aims of OSPAR's Hazardous Substances Strategy is that concentrations of naturally occurring chemicals should approach background concentrations, and concentrations of man-made chemicals should be zero. Progress on the implementation of this strategy is monitored through the Joint Monitoring and Assessment Programme (JAMP) of chemicals for priority action and hazardous substances in general. The main objectives of the JAMP for the period 2010–2014, which seek to support the implementation of the OSPAR strategies and the EU MSFD are:

1. the continued implementation and development of existing OSPAR monitoring programmes and, where necessary, the development of additional coordinated monitoring programmes to take account of criteria, methodological standards and indicators for good environmental status, and the pressures and impacts of human activities;
2. development of tools for the delivery of integrated environmental assessments of the OSPAR maritime area or its regions, linking human activities, their pressures, the state of the marine environment, and management responses. Where relevant, these tools should support the exploration of new and emerging problems in the marine environment;
3. the preparation of integrated environmental assessments of the implementation of the OSPAR strategies, including in particular the assessment of the effects of relevant measures on the improvement of the quality of the marine environment. Such assessments will provide additional information and assessments in respect of the MSFD, enhance the OSPAR quality status reports (QSRs), take into account the Directive's obligations for regional cooperation, and help inform the debate on the development of further measures.

Aqueous inputs (direct or riverine) of contaminants, together with atmospheric deposition, are important sources of contaminants to OSPAR marine waters. Dynamic equilibria exist between the dissolved fractions of the total burden of contaminants, such that contaminants are partitioned between the dissolved state and particulate and colloidal phases in the water column, as well as becoming associated with bottom sediments and biota. The rates of exchange of contaminants between the water and the sediment or biota mean that changes in inputs are likely to be reflected more rapidly in the water than in, for example, bottom sediments. However, this sensitivity to change, and the partitioning between components of the aqueous phase, are also reflected in relatively high spatial and temporal variances in the observed concentrations. The selection of water as a monitoring matrix can therefore be appropriate for a number of reasons. These include the ability to observe short-term variations in contaminant pressure on organisms. Focusing on contaminants that partition strongly into the water rather than the sediment or biota can lead to water being the preferred

matrix for monitoring. OSPAR background documents on chemicals for priory action may provide valuable information with regard to the preferred monitoring matrix. In the context of the JAMP, coordinated monitoring of contaminants in seawater may be carried out in relation to the temporal changes in the degree of pollution, its spatial variation, or as an element of integrated monitoring and assessment of contaminants and biological effects.

Temporal trend monitoring can assess the effectiveness of measures taken to reduce contamination of the marine environment. The statistical assessment of a trend over a longer period also supplies a more reliable assessment for the environmental status within a certain period. The fitted value of the last year measured has been used in OSPAR CEMP assessments as the optimum value for comparing against assessment criteria and hence for assessment of the actual environmental status. In such a way, the within- and between-year variability is taken into account.

Spatial distribution monitoring can describe the existing level of marine contamination widely through the convention area. The measured levels can be compared to background or close to background concentrations, as well as to levels describing thresholds below which no chronic effects are expected to occur in marine species, i.e. environmental assessment criteria (OSPAR, 2009).

Contaminant analysis of seawater can be an element of integrated monitoring and assessment, where chemical and biological effects measurements are combined, in order to assess potential harm to living resources and marine life (OSPAR, 2012). The role of chemical measurements in integrated chemical and biological effects monitoring programmes is to support biological effects programmes by providing information to help identify the chemical causes of observed biological effects. In general, chemical measurements in seawater should contribute to improve and extend OSPAR's monitoring framework and better link it with the understanding of biological effects and ecological impacts of individual substances and the cumulative impacts of mixtures of substances.

Furthermore, beyond the objectives of the JAMP, monitoring of contaminants in water can provide information on the fate of contaminants in the environment, e.g. transformation, partitioning, and transport processes.

3. Quantitative objectives

Seawater monitoring should provide concentrations of target analytes in water, which are representative of the location and time of sampling. General considerations regarding the specification of quantitative objectives for monitoring are given in the JAMP (OSPAR, 2010). More specifically, the following issues should be considered prior to water monitoring: contaminant speciation, detection limits, detectability of temporal and spatial trends, and costs.

3.1. Contaminant speciation

Trace metals and organic contaminants can exist as freely dissolved species in water or bound to colloids and suspended particulate matter (SPM). Trace metals can also exist as inorganic and organic complexes. The targeted contaminant fraction determines which sampling and/or pre-treatment method to use:

- Analysis of unfiltered water samples yields the sum of the concentrations of contaminants that are freely dissolved, complexed, and bound to colloids and SPM. These samples are also referred to as total water or whole water samples.
- Filtered water samples can yield the concentrations in SPM (by analysis of the residue on the filter) and the concentrations of contaminants that are freely dissolved, complexed, and bound to colloids (filtrate). However, many organic contaminants are known to exchange freely between dissolved and other phases in the water. The removal of components of the particulate matter is very likely to alter the position of these equilibria, while the introduction of filter material, container walls, etc. provides additional phases taking part in the equilibration processes. The complete separation of dissolved, colloidal, particulate matter is therefore a difficult task.
- Passive sampling yields the concentrations of freely dissolved contaminants (organics) or freely dissolved and complexed contaminants (trace metals).

The choice of the targeted contaminant fraction may be pre-defined by legal obligations. For example, monitoring under the Water Framework Directive requires the monitoring of metal concentrations in filtered water, and of organic contaminants in total (i.e. unfiltered) water.

3.2. Detection limits

The sample size has to be sufficient to support the desired detection limits for the contaminants of interest, for example to enable descriptions of spatial and temporal trends. For example, one litre discrete water samples may be sufficient for time trend monitoring of PAHs in contaminated harbours, but may be insufficient for monitoring programmes in open waters. For consistency with Commission Directive 2009/90/EC, a limit of quantification (LOQ) should be equal to or below a value of 30% of the relevant assessment criterion, e.g. the Environmental Quality Standard.

3.3. Statistical significance and power

In the context of temporal trend monitoring, it is important to know the statistical power of a time-series to detect changes, i.e. the probability of detecting true trends in concentration in the presence of variance associated with sampling, analysis, and field variability. The necessary or possible power of a monitoring programme will vary with the contaminant and area being investigated. One approach would be to estimate the power of the time series based on the "random" between-year variation. Alternatively, the lowest detectable trend could be estimated at a fixed power. A quantifiable objective could be to detect an annual change (dC/dt) of 5% within a time period of 6 years with a power of 90% at a significance level (α) of 5%. In the case of an expected decrease, the null hypothesis would be chosen as $dC/dt=0$ and the alternative hypothesis as $dC/dt< 0$.

A spatial monitoring programme should enable Contracting Parties to describe the distribution of contaminant concentrations in the survey area, for example to draw maps. These data can provide information to assist in the identification of representative stations for temporal trend studies, or for refinement of spatial surveys, and to implement measures where considered necessary. Statistical procedures can be used to estimate the number of samples and sampling sites needed to meet the required confidence level (i.e. to avoid Type I errors) and statistical power (to avoid Type II errors).

3.4. Costs

The concentrations of contaminants in water, as determined by discrete sampling, are commonly found to be quite variable, both in space and time, and meeting ambitious quantitative objectives may require extensive replication. Seawater sampling for contaminant analysis often requires equipment that is expensive to buy and maintain in good condition to keep the process blanks at low levels. The need for, and cost, of replicate water samples should be carefully considered in determining achievable quantitative objectives for a water-based monitoring programme. Therefore, it is often necessary to balance the scope and performance of monitoring programmes with available budgets.

4. Sampling strategy

The sampling strategy should reflect the purpose of the monitoring programme according to the JAMP (OSPAR, 2010) in relation to the OSPAR Hazardous Substances Strategy. Where applicable, the sampling strategy should consider requirements of the EU WFD (EU, 2000) and MSFD (EU, 2008); in all cases the quantitative objectives of the monitoring programme should be met (see Section 3). In accordance with the JAMP Guideline on Integrated Monitoring of Contaminants and Their Effects, seawater sampling should be carried out at the same time and locations as the sampling of other matrices (sediment, biota) and biological effects measurements (OSPAR, 2012).

A coherent approach to the detailed definition of a sampling strategy should take into account knowledge of the physical and biological oceanography of the area and requires consideration of temporal sources of field variance, such as seasonal factors, and spatial factors, such as the changes in location and water depth within the survey area. The analyte in question (its physical-chemical characteristics and expected concentration), as well as environmental conditions and practicalities, will further determine how samples are taken, e.g. what equipment is used and what volumes are required. However, sampling strategies also include compromises between scientifically advisable approaches and the economical and logistical frames of the sampling effort (see Section 3). It is therefore important that the objectives of monitoring programmes are expressed in quantitative terms and that they are achievable.

4.1. Temporal trend monitoring

The ability of a programme to identify temporal trends strongly depends on the extent to which unwanted sources of variability can be controlled. The short-term (< 1 year) temporal variability of contaminant concentrations in water is potentially very large. Concentrations may be subject to day-night variations in input and removal processes (Jaward *et al.*, 2004). In addition, concentrations at a fixed geographical position may vary over the tidal cycle (e.g. in estuaries). Further temporal variability may arise from variation in local inputs, such as discharges from ships, seasonality in the riverine discharge, changes in atmospheric deposition during rainfall events, and seasonal differences in seawater stratification. Some measures can be taken to reduce short-term temporal variability. These include sampling at pre-defined times of the year and at the same phase of the tidal cycle (e.g. always at high tide), although for ship-based discrete sampling it should be recognized that logistic constraints do not always allow such measures to be taken.

4.2. Spatial distribution monitoring

Analyte concentrations in seawater will vary between locations and with water depth, due to various physical and biogeochemical processes and the distribution of inputs. The expected spatial variability is an important factor in the development of an adequate geographical sampling scheme, i.e. the outline of the station grid and its vertical resolution (Brügman and Kremling, 1999). It should be recognized that the identification of spatial patterns may be obscured by

temporal variability (see Section 3.1), and that the same measures to reduce this source of variability also apply here. If the aim of the programme is to identify local sources of contaminants, then the sampling grid should be denser in the vicinity of suspected sources. Often, the variability of salinity or SPM content of the water can give an indication of the variability of pollutants and may even act as "normalization" factors.

4.3. Sampling method considerations

The proportion of the total concentration of a contaminant which is freely dissolved in the water phase increases with polarity of the pollutants (see Section 3). On the other hand, non-polar pollutants sorb to SPM and sediments and are thereby removed from the water column by sedimentation. For these contaminants, additional factors that should be taken into account are the SPM content and the volume of water that is sampled (see Section 3). These factors are important in filtration-extraction methods because the particle-bound and colloidally bound contaminant fractions that escape phase separation depend on the extent of filter clogging (Hermans *et al*., 1992). The measurement of SPM concentrations is even more important for monitoring contaminants in total water. The required water volume should be estimated before the sampling campaign, taking into account the method detection limits (see Section 3).

4.4. Supporting data

It is important that as much information as possible is collected concerning the waterbody being sampled. This includes co-factors such as salinity, SPM concentrations, and temperature. Whenever possible, sampling should be done as part of an integrated monitoring programme that includes the measurement of biological effects. These data should be obtained at the same time and locations as sampling for contaminant analysis.

4.5. Statistical considerations

Prior to starting a full-scale monitoring study, the available information on temporal variability should be carefully evaluated, possibly amended by a small-scale pilot programme. This evaluation should include a statistical assessment certifying that the objectives of the monitoring study can be met (see Section 3).

If no previous information exists, the sampling strategy can be based on a combination of general statistical principles and expert knowledge about sources and fate of the studied substances in the investigated sea basin. The statistical approach could include the principles of stratified sampling: First, the sampling area under consideration is partitioned into smaller more homogeneous areas, so-called strata. This can be based on simple information, such as depth, distance to land, or measured or modelled salinity. A successful stratification is characterized by a small variation of the measured concentrations within each stratum and a substantial variation between strata. For optimal allocation of the samples, the size (volume or area) of each stratum should be determined. Assuming that there are m strata with volumes $V_1, \ldots, V_m$ and that the standard deviation of the target variable is about the same in all strata, the number of samples n_j in stratum j shall be taken approximately proportional to the volume V_j, i.e.

$$n_j \approx n\frac{V_j}{V}$$

where V is the total volume of the investigated sea basin and n is the total number of samples.

If the standard deviation of the target variable varies from stratum to stratum, more samples should be taken in strata with high standard deviation. More specifically, the sample numbers chosen should aim at making n_j proportional to S_jV_j, where S_j is the standard deviation in the jth stratum, i.e. letting

$$n_j \approx n\frac{S_jV_j}{\sum_{j=1}^{m} S_jV_j}$$

Finally, the average concentration in the study area is estimated to be

$$\sum_{j=1}^{m} V_j\overline{X}_j / V$$

where $\overline{X}_j$ is the average observed concentration in the jth stratum.

4.6. Discrete sampling versus time-integrated sampling

Concentrations of contaminants in water respond quickly to changes in inputs and other environmental conditions, unlike concentrations in sediments and biota. This low level of time integration can be of advantage in detecting peak events but, on the other hand, concentrations in water are likely to show relatively high variability, which can have drawbacks in long-term monitoring and may require high sampling frequencies, causing high costs.

The influence of temporal variability may be reduced by time-integrated sampling. However, continuous water intake over a prolonged time period, followed by filtration and extraction, may often prove to be impractical and costly, particularly for ship-based sampling programmes. Unattended integrative devices, such as passive samplers (PSDs) also yield a time-integrated concentration if the necessary calibration parameters are available for the target analytes. Considerations for evaluating whether the necessary PSD calibration parameters are available for non-polar organic analytes are given by Lohmann *et al.* (2012). PSDs for polar contaminants (pharmaceuticals, detergents, and personal care products) are insufficiently mature for quantitative spatial and temporal trend monitoring at present, but may be useful in initial surveys. Diffusive gradients in thin films (DGT) is a mature PSD technique for trace metals, but its application in the marine environment has been quite limited so far (Mills *et al.*, 2011). All PSDs require suitable deployment sites, such as jetties, buoys, bottom landers, long-term moorings, etc, which always have to be visited twice and some losses due to other marine activities may be expected. If the monitoring programme requires sampling of total water, this will limit the applicability of PSDs.

5. Sampling equipment

The choice of sampling equipment depends on the physical-chemical properties and expected concentrations of the analytes, on the depth and location of the sampling site, and on the available infrastructure. All materials used for the sampling equipment (sample containers, tubing, connectors, valves, pumps, filters) should neither absorb nor release the target analytes, or any non-target substance that interferes with the chemical analysis. Contaminants are held in a range of dissolved, colloid, and particulate phases. These have a potential to interact differently with sampling equipment, and also for contaminants to exchange between phases during sample processing. Sampling equipment and processing therefore needs to be rigorously tested before adoption in large-scale monitoring programmes.

Since concentrations of organic contaminants and metals in seawater are usually very low, large volumes of water must be sampled. Contamination of the sample by compounds that leach out of the sampling equipment as well as analyte loss due to wall sorption are serious issues which may affect the integrity of seawater samples.

Sample contamination from the atmosphere should be avoided (e.g. paint and rust particles, engine exhausts, atmospheric background). To minimize contamination from the atmosphere, the surfaces of the sampling equipment in contact with the sample should be isolated from the atmosphere before and after the sampling, including storage of the equipment. These surfaces should be cleaned using appropriate solvents prior to sampling. Equipment blanks and recovery samples yield important quality control information that can be used to assess sample contamination and analyte losses, bearing in mind the potentially site-specific nature of airborne contamination.

Concentrations of target analytes in the water may be elevated because of leaching from the sampling platform itself (e.g. polyaromatic hydrocarbons (PAHs), organotin, polychlorinated biphenyls (PCBs), iron, and chlorofluoroalkanes can be released from the ship during ship-based sampling). The ship's keel should be at an angle of 20 to 40 degrees to any current coming from the bow at the sampling side (typically starboard side), to minimize any influence from the ship's hull.

Since the sampling equipment passes through the air-water interface, contamination from the sea surface microlayer is a significant risk. Concentrations of dissolved and particulate matter are elevated in this microlayer, and the associated analytes may therefore contaminate samples that are taken at larger depth. Sample contamination from the microlayer can be avoided by closing the sampling equipment during passage through the sea surface and only allowing sample intake at the intended depth.

5.1. Trace metals (including MeHg)

Contamination from the ship has to be avoided at all times. For analyses of trace metals, all contact between the seawater sample and metal must be avoided. On approaching a station, the sampling for trace metals has to be performed immediately. Hydrographical information about water depth and the stratification of the water column should be available.

Discrete samplers that are specially designed for trace metal analysis should be used, e.g. GO-FLO (from General Oceanic), available in sizes from 1.7 to 100 litres, or MERCOS samplers (from Hydrobios; or modified version, size 0.5 litre). They are typically operated on a Teflon, polymer, or Kevlar jacketed stainless steel hydrographic wire, tensioned

by a coated bottom weight. The messengers should also be free of metals; any essential metal parts should be of seawater resistant stainless steel (V4A).

Samples should be taken so as to avoid contamination by leachate from the hull of the ship. Sampling bottles should be made of plastic with low metal content, e.g. special low-density polyethylene (LDPE) bottles. For mercury, glass should be preferred if the samples are stored for a longer period. Teflon bottles may also be used, but they are relatively expensive and, depending on the manufacturing process, may have a relatively rough inner surface.

Pumping using metal-free devices may be an alternative to discrete sampling, e.g. for separating SPM by subsequent centrifugation, but is not preferable when sampling from a ship at distinct sampling depths or in the open sea where concentrations are very low. More details on sampler types are described in the Technical Annex.

After sampling, the sampler should be placed immediately in a plastic bag or box or an aluminium container (if aluminium is not determined), followed by transport to a clean-room or laboratory with a clean-air bench. These measures are particularly critical for open sea samples where the expected concentrations of trace metals are very low.

5.2. Organic contaminants

Concentrations of organic contaminants in seawater are usually very low. In order to reach the projected LOQs in the low pg l^{-1} range, large water volumes (10 to 100 l or more) have to be collected and extracted. With modern analytical equipment, these LOQs are often not limited by the signal intensity in the instrumental analysis, but by blank levels and interferences from the matrix background.

Hydrophobic compounds occur in a continuum of dissolved, colloidal, and particulate-bound forms. Unless a total concentration is to be determined, the compound partitioning must not be altered during sampling and subsequent treatment. This is very challenging, as the separation process must be contamination-free and should not change the concentration distribution. It should be applied during or immediately after sampling. For details, see Section 6.2.

Sometimes blank problems can only be overcome by increasing the sample size. However, the maximum sample size may be limited by operational constraints, such as container size for discrete samplers, pumping time, and the ability to process large water volumes. Blank levels can be reduced by minimizing the size of the sampling equipment (e.g. short inlet tubes) and by using sampler designs and handling procedures that minimize exposure to the atmosphere (short assembly/disassembly times). The use of *in situ* filtration/extraction equipment that is both compact and easy to operate combines the advantages of small size and short exposure to the atmosphere. This holds even stronger for passive samplers (see Section 4.6), provided that the sampling phase is sufficiently clean and that times of exposure to the atmosphere during deployment and retrieval are sufficiently short.

The materials used for the sampling equipment depend on the target contaminants. Sampling equipment for organic contaminants in seawater is preferably made of glass or stainless steel. Teflon parts are often used for legacy persistent organic pollutants (POPs), while they cannot be used for sampling of fluorinated compounds. Before use, the equipment has to be cleaned, e.g. rinsed with appropriate organic solvents. Examples of sampling equipment suitable for organic contaminants are presented in the Technical Annex.

6. Storage and pre-treatment of samples

The storage and pre-treatment of samples should be carried out in full awareness of the risks of contamination or analyte loss if samples are handled incorrectly. Appropriate measures should be taken to avoid contamination, such as wearing clean gloves, pre-cleaning equipment, etc. All storage and pre-treatment steps should be fully documented for each sample. Field control samples (for assessing sample contamination) and surrogate spikes (for assessing analyte losses) should be processed regularly as part of the quality assurance and control procedures (see Section 8). All storage and pre-treatment steps should be fully validated prior to the start of a monitoring programme.

6.1. Storage

It is advisable to process samples as soon as possible rather than store them for a longer period of time. Storage of samples increases the risk of changing concentrations, by microbial degradation or sorption processes. However, appropriate laboratory facilities for handling of samples for trace analyses need to be available. If this is not the case, samples may have to be conserved. Water samples for metal analysis are typically acidified for conservation purposes. Sub-sampling of seawater, if required, should preferably be performed immediately after sampling.

Water samples for organic pollutants generally are impractical to store because of their large volumes. Instead, they are extracted onboard by liquid–liquid extraction (LLE) or solid-phase extraction (SPE) and the extracts or adsorbent cartridges are stored under cool (< 4°C) and dark conditions. If water samples must be stored, this should also be in the

dark and in a refrigerator (4°C). Preferably, internal standards (e.g. isotopically labelled analogues) should be added before extraction or/and storage. Storage times should be kept as short as possible and the stability of all compounds during storage must be checked.

Only appropriate (pre-cleaned) containers should be used for short- or long-term storage. The analytes of interest determine the appropriate container material (plastic, glass, metal), the need for acidification, and the optimal storage temperature. All storage conditions should be fully validated by the laboratory that carries out the monitoring, since sample contamination and loss of analyte may be affected by subtle changes in the materials and procedures for sample storage. SPM samples should always be stored frozen until further analysis.

6.2. Sample pre-treatment

The need for filtration of samples is mainly determined by the monitoring programme which typically will specify the analysis of either filtered or unfiltered water (total water, whole water). No pre-treatment is required for the analysis of whole water, although acidification may be necessary as part of the extraction procedure, depending on the analyte and on the extraction method used.

Filtration is the preferred technique to separate the dissolved phase from the SPM for small volume samples (e.g. for metal analysis). Polycarbonate or cellulose acetate filters with a pore size of 0.45 µm are frequently used for trace metal determinations, whereas glass fibre filters (0.7 µm or 1.2 µm pore size) are commonly used in the analysis of non-polar and polar organic contaminants. The efficiency of the separation between dissolved and particulate contaminants depends on the pore size of the filters, and may also depend on SPM content of the water and on the sample intake (see Section 4). Adsorption of dissolved analytes to the filter may be an issue for some compounds, and should be addressed during method validation.

A flow-through centrifuge is suitable for obtaining SPM from large volume samples, but less suitable for obtaining particle free water as the separation is incomplete. In general, the efficiency of the separation depends on the geometry and operating conditions of the centrifugation equipment (residence time, effective gravity force), as well as on the density and size of the SPM. Filtration is more effective in this respect, but also more susceptible to artefacts and more time consuming. Ideally, filtration should occur online while sampling or immediately after sampling.

7. Analytical procedures

Analytical methods should be specific to the target analytes and sufficiently sensitive to allow analyses of seawater samples which generally have low concentrations of contaminants. They should meet minimum performance criteria consistent with Commission Directive 2009/90/EC, including an uncertainty on measurements < 50%, estimated at the level of the relevant Environmental Quality Standard, and an LOQ ≤ 30% of the Environmental Quality Standard. If no method meets the minimal performance criteria, the best available analytical method, not entailing excessive costs, should be used. All analytical methods should be capable of being brought under statistical control to ensure adequate quality assurance and quality control. It should be noted that analyses at such low concentrations require extensive experience.

7.1. Trace metals

Analysis of trace metals in seawater generally includes pre-treatment and pre-concentration steps, followed by detection using element-specific spectrometric instrumental procedures, e.g. graphite furnace atomic absorption spectrometry (GFAAS), inductively coupled plasma mass spectrometry (ICP–MS), anodic stripping voltammetry (ASV), and total reflection x-ray fluorescence (TXRF). For mercury, further methods and instruments are used, such as cold vapour atomic absorption spectrometry (CVAAS) and cold vapour atomic fluorescence spectrometry (CVAFS). These techniques are usually combined with a pre-concentration by amalgamation. ICP–MS is also used for mercury analysis.

7.2. Organic contaminants

Organic contaminants are usually found in the water phase at low concentrations, entailing the need for an extraction and enrichment step (e.g. SPE, LLE, solid-phase micro extraction (SPME)) and a selective chromatographic/detection step (e.g. GC–$MS^{(n)}$, GC–ECD, LC–$MS^{(n)}$, LC–Fl.) within every analytical procedure. Depending on the analytes chosen, the water body studied and expected pollutant concentration, clean-up may be necessary. Although GC–MS/MS and HPLC–MS/MS are very selective techniques, it is good practice to use a second MS transition as a qualifier.

8. Quality assurance (QA)

The quality assurance programme should ensure that the data conform to the quantitative objectives of the programme (see Section 3). The laboratory must establish a quality assurance / quality control system, if necessary consistent with

requirements in Commission Directive 2009/90/EC. All field and laboratory procedures should be fully validated, and the laboratory should also participate in intercalibration exercises and proficiency testing to provide external verification of results. The quality assurance procedures should cover sampling design, sampling, sample storage, analytical procedures (including field controls, analytical blanks, and recoveries), equipment maintenance and handling, training of personnel, data management, and an audit trail.

The use of a second (and different) sampling method, carried out simultaneously to the routine procedure, can be included in the validation process. All QA and QC data should be fully documented.

Because of the extremely low concentrations of pollutants in seawater, blank problems are generally more relevant and more difficult to control than in other matrices. Even ultra-pure chemicals and solvents used sometimes have to be purified before use. Concentrations are often close to the LOQs, which means difficult calibration and integration, and reduced analytical precision.

In addition, the following problems are encountered specifically in seawater analyses of organic contaminants:

- Because of the large sample volumes, it is not possible to analyze replicate samples on a routine basis or to take samples for back-up analysis. However, it is often possible to make a plausibility check by comparing the results with those of samples taken from adjacent stations in a homogeneous water body. Homogeneity can be assessed from oceanographic parameters, like salinity.
- No certified reference materials are available for organic contaminants in seawater. Therefore, laboratory reference materials have to be used, which should preferably be a natural or spiked extract from a typical monitoring station. Extraction efficiencies should be checked by standard addition tests.
- Laboratory performance studies (e.g. by QUASIMEME) are difficult to perform and to evaluate because sample volumes in these studies (max. 1 l) differ from those used in real analysis (>10 l). Thus, concentration ranges in the tests are often higher than in real-life samples.

For temporal trend monitoring in particular, it is extremely important to perform reliable and reproducible high-quality analyses over decades. Therefore, such analyses require well-documented procedures and experienced analysts (see Section 7).

9. Reporting requirements

Secure data storage and appropriate access to the data should be ensured by submission of data to national databases and to the ICES database. Reporting requirements will depend on the database. For entry of OSPAR data into the ICES database, data of trace metals and organic contaminants should be reported in accordance with the latest ICES reporting formats.

The calculation of results and the reporting of data can be major sources of error. Control procedures should be established in order to ensure that data are correct and to avoid transcription errors. This could include comparisons with independently obtained results for the same area or with typical concentration intervals. Data stored in databases should be checked and validated, and checks are also necessary when data are transferred between databases.

Concentrations of trace metals and organic contaminants in seawater should be given in weight per volume (e.g. ng l^{-1}). To ensure correct interpretation, reporting should include information on the sampling method, filtration (filter type and pore size), storage/conservation, and analytical method. Minimum performance criteria such as LOQ and uncertainty measurement along with relevant QA/QC data such as reference material analyses should be included in the report.

The purpose of the monitoring, geographical coordinates, and the name of the sampling stations should be reported in the data as well as being defined in the OSPAR Station Dictionary (http://www.ices.dk/datacentre/accessions/). Sample depth, suspended particulate matter concentration, and physicochemical parameters at the time of sampling, such as air and water temperatures, salinity, pH, and weather conditions, should also be reported.

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Technical Annex: Sampling equipment for analysis of trace metals and organic contaminants in seawater

1. Trace metals

1.1 Discrete sampling

An example of a discrete sampler is the GO-FLO sampler by General Oceanics (Figure 1). This sampler consists of a cylinder with an inner Teflon-coating which can be closed and lowered into the water column and opens automatically at a certain depth (ca. 10 m) by hydrostatic pressure. This avoids contact of the sample with the water surface where some contaminants can accumulate. At the desired depth, a messenger is sent on the hydrographic wire (made of Teflon coated stainless steel, polymer, or preferably Kevlar) to release the closing valves in both ends of the sampler. Each bottle can be equipped with a second messenger that is released when the valves close. Water samples can be collected from a range of depths by mounting a series of bottles along the cable.

A variety of the GO-FLO sampler is the reversing water sampler. The messenger releases the sampler from the upper attachment, it rotates, and closes the two valves. If a special thermometer type is attached to the sampler, it fixes the actual temperature at the sampling depth, which can be determined later on board. This accessory can be used when no CTD-sensor is used to record the temperature profile.

Generally, all samplers must be cleaned before the first use by rinsing the inner surfaces with diluted hydrochloric acid. In the open sea, this may not be necessary between sampling where rinsing with deionised water is sufficient in most cases. In the open sea, seawater is sufficiently clean to rinse the outer surface. Samplers with rubber parts which cannot be acid-cleaned or cannot be closed during deployment should be avoided.

Figure 1 Picture of a GO-FLO sampler (General Oceanics; photo courtesy of IFREMER, France).

The MERCOS sampler (Hydrobios Kiel) is designed for two 500 ml thick-walled cylindrical or ball-shaped Teflon bottles, which are closed by two silicone tubes of different diameters in the water. As the bottles are filled with air, the operating depth is restricted to about 50 m for the cylindrical and about 200 m for the globular type. However, this sampler is no longer offered by the manufacturer (http://www.hydrobios.de, 2012).

A modified version for four bottles was developed by the Bundesamt für Seeschifffahrt und Hydrographie (BSH, Germany), maintaining the triggering device, but using LDPE bottles of low metal content material (NALGENE) that are protected against the water pressure by a polyacrylate mantle. The LDPE bottles are cheaper and easier to clean due to the smooth inner surface compared to the relatively rough texture of the thick-walled Teflon bottles. Therefore, the LDPE usually show much lower blank values.

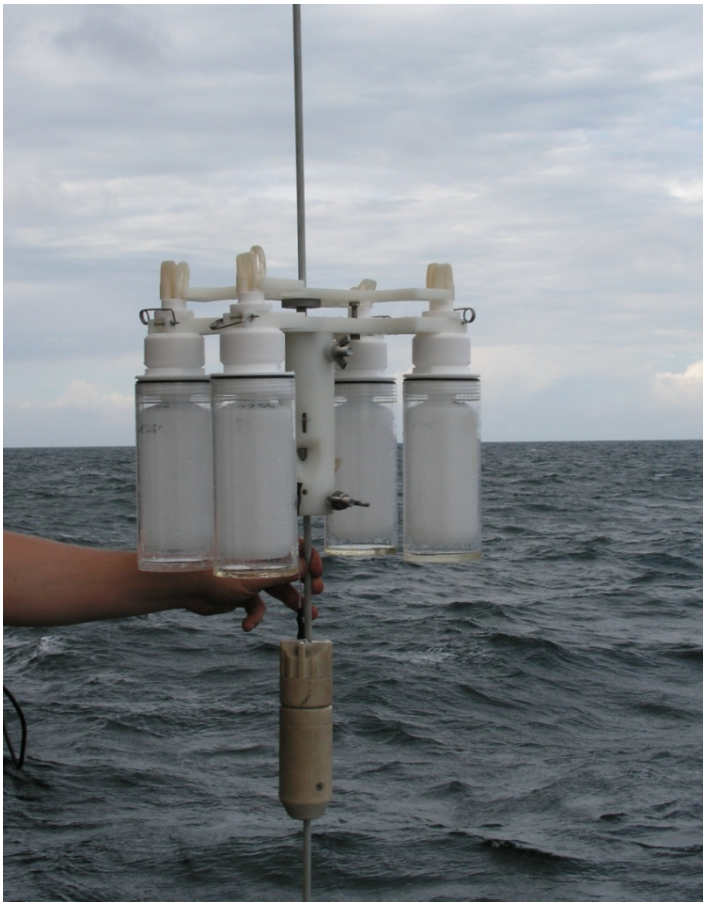


Figure 2 **Modified MERCOS water sampler of the second generation for four bottles, manufactured by BSH, Germany (photo courtesy of S. Schmolke, BSH, Germany).**

1.2 Sampling by pumping

For depths down to 100 m, perhaps even 200 m, it can be practicable to pump seawater up through silicone or Teflon tubing, optionally including in-line filtration. The tubing should be cleaned by pumping acid (e.g. 10% hydrochloric acid) prior to sampling. The first litres of seawater sampled should be subsequently discarded. A peristaltic pump or Teflon piston pumps are suitable. The peristaltic pump can be placed between the sampling tube and the filter. The outflow from the in-line filter can then be collected in polyethylene bottles, Teflon bottles, or in glass or quartz bottles for mercury analyses.

2. Organic contaminants

Large volumes of seawater samples are usually needed for the analysis of organic contaminants. Sampling devices depend on the amount of sample to be processed and the method of extraction (liquid–liquid extraction (LLE) or solid-phase extraction (SPE)).

LLE and SPE do not yield exactly the same concentrations as they use different extraction principles. While SPE effectively extracts only freely dissolved compounds, LLE extracts freely dissolved compounds and also compounds complexed with humic acids and, in part, compounds bound to particles (Sturm *et al.*, 1998). Non-polar compounds can be extracted by either LLE or SPE, whereas the extraction of polar compounds generally requires SPE.

Volumes of 1 to 100 l can be sampled by discrete sampling and/or pumping and are usually extracted either by LLE or SPE. Sample volumes >100 l are generally sampled by pumping and extracted by SPE.

2.1 Discrete sampling

Several different sampling devices have been designed for discrete sampling depending on the volumes needed and the extraction techniques to be applied.

All-glass bottle samplers for volumes of 10 L and 100 L are shown in Figure 3. They are mounted in a stainless steel cage and lowered on a hydrographic wire down to the desired sampling depth and opened under water. After filling, the sampler is brought on deck of the ship and the sample can be extracted by LLE directly in the sampler (using a non-polar solvent) or by SPE. For example, non-polar pollutants like organohalogen pesticides (e.g. DDx, HCH, HCB, dieldrin, endrin) can be extracted and enriched from seawater by means of LLE using hexane or pentane.

Gaul and Ziebarth (1983) described a 10 l glass sampler allowing extraction in the sampling flask itself, thereby minimizing uncertainties arising from sample handling, blanks, adsorption, etc. Later, the same principle was expanded to a 100 l flask, thus increasing the sample volume and lowering the limit of quantification (LOQ) by a factor of 10 (Theobald *et al.*, 1990). Figure 3 shows pictures of 10 l and 100 l sampling bowls. Extraction is done by agitating the samplers with 0.2 and 1 liter of pentane, respectively, using a stirrer. The glass sampler can be used to a depth of 2000 m (10 l) and 100 m (100 l).

Collecting samples at greater depth can be done with stainless steel bottles (Figure 4) holding about 30 litres. This type of sampler was developed based on experience with Niskin and Go-Flo type bottles, and has been used in analyzing dissolved herbicides in water samples collected down to 3000 m depth.

Figure 3 **Left: BSH all-glass bottle water sampler (10 l). Right: 100 l glass flask sampler for sampling seawater for the analysis of organic contaminants.**

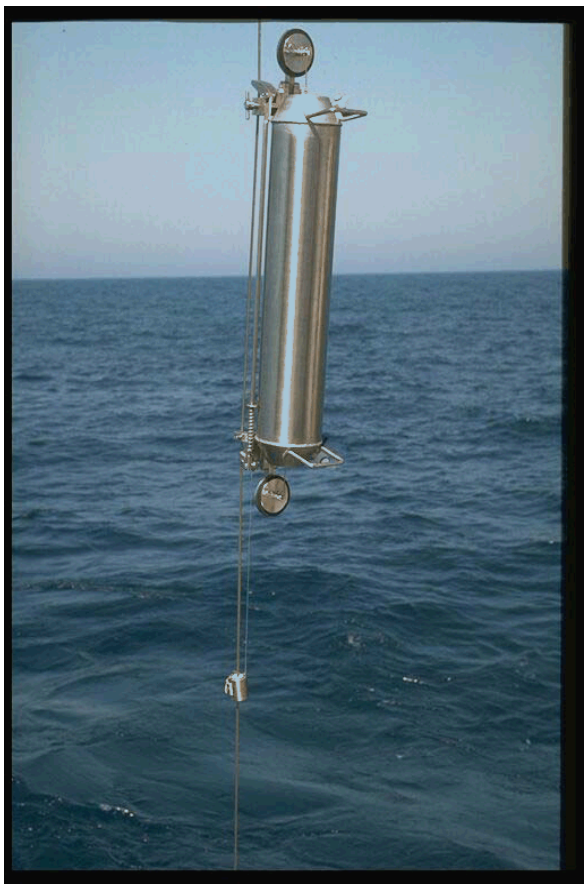


Figure 4 **A stainless steel sampling bottle, for subsequent analysis of organic contaminants in seawater.**

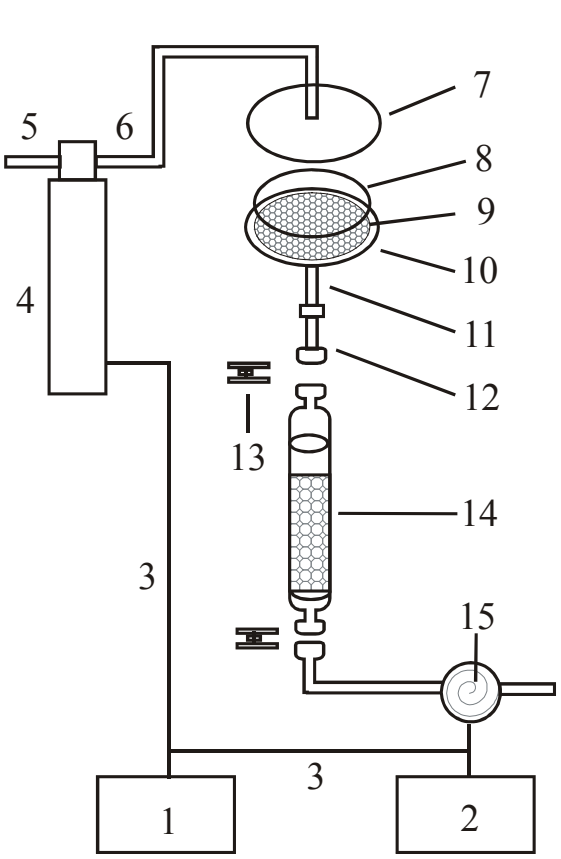
2.2 Sampling by pumping – *In situ* filtration and extraction

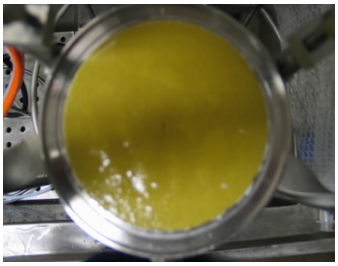
For larger volumes of 200 to 1000 l, Schulz-Bull *et al.* (1995) described an SPE procedure using large extraction cartridges filled with XAD resins. With this adsorbent, they obtained good extraction recoveries for PCBs, DDT, and PAHs, but not for HCH.

Sampling by pumping can be performed with compressed air Teflon pumps (not suitable for subsequent analysis of perfluorinated compounds). In order to equilibrate the system with the sampling water, the water is pumped for about ten minutes before the actual sampling begins. Then the sampling bottles are thoroughly rinsed with the sample, before beginning the sampling itself. The hose is kept away from the ship's hull while the system is being rinsed, and during the collection of the sub-surface samples.

In situ filtration and solid-phase extraction sampling devices may minimize the risk of sample contamination during sampling. A typical *in situ* pump system, the Kiel In-Situ Pump (KISP), has been widely applied to the extraction of organic contaminants in seawater (Petrick *et al.*, 1996). A modified KISP has been described for seawater sampling on-board research vessels (Ebinghaus and Xie, 2006). Briefly, as shown in Figure 5, KISP includes a filter holder, a polymeric resin column, a pump, and a flowmeter. A glass fibre filter (pore size 0.7 µm) is used to recover the particulate phase and a glass column packed with polymeric resin for the dissolved phase. The KISP can be easily operated on board by connecting it to the ship's seawater intake system for sampling seawater at certain depths. The pump system assembly with batteries can be deployed at different depths on a hydrographic wire, and the pumping can be started and ended by remote control.

The original KISP contains some plastic parts and connections, which may present a contamination risk for some organic contaminants, such as brominated flame retardants, alkylphenols, and plasticizers. Low blanks and detection limits have been obtained from KISP samples for legacy persistent organic pollutants (POPs), such as PCBs, DDTs, and HCHs (Lakaschus *et al.*, 2002; Sobek and Gustafsson, 2004). However, it is recommended that these parts are replaced by stainless steel or glass if KISP is to be applied for sampling seawater for the determination of other organic contaminants. Surrogate standards can be added to the resin column before sampling to control the extraction recoveries and storage. It should be noted that the validation of the *in situ* pump sampling method is difficult, and extraction efficiency may depend on dissolved organic matter and humic substances.





GF filter after sampling of 300 l seawater

Figure 5 **Schematic presentation of the Kiel In-Situ Pump (KISP). 1: flowmeter controller; 2: flowmeter; 3: cable connections; 4: pump; 5: pump inlet; 6: pump outlet; 7: stainless steel deck of filter holder; 8: GF 52 filter; 9: glass plate; 10: filter holder; 11: stainless steel tubing; 12 glass connect; 13 adjustable clip; 14: resins column; 15: counter of flow meter.**

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Annex II:

HELCOM (2012a). Manual for marine monitoring in the COMBINE programme. Annex B-11, Appendix 1. Technical Note on the determination of trace metals (Cd, Pb, Cu, Co, Zn, Ni, Fe) including mercury in seawater (3.1.2)

HELCOM Manual for marine monitoring in the COMBINE programme

ANNEX B-11, APPENDIX 1. TECHNICAL NOTE ON THE DETERMINATION OF TRACEMETALS (CD, PB, CU, CO, ZN, NI, FE), INCLUDING MERCURY, IN SEAWATER

Introduction

General techniques which address the questions of water sampling, storage, filtration procedures and determination of trace metals in natural sea water are described by Sturgeon and Berman (1987) and Gill and Fitzgerald (1985, 1987).

For the determination of mercury in sea water, the chemical species of this element are of importance. Therefore, a differentiation between the several Hg species, including ionic, volatile, dissolved (organic) complexes or particulate adsorbed Hg, has to be considered during sample preparation.

Several definitions of mercury compounds are common (Cossa et al., 1996, 1997), for example:

- Reactive mercury (HgR): A methodologically defined fraction consisting mostly of inorganic Hg(II).
- Total mercury (HgT): Mercury content of an unfiltered sample, after digestion with an oxidizing compound (e.g., K MnO4).
- Total dissolved mercury: Mercury content of a filtered sample, after digestion with an oxidizing compound (e.g., K MnO4).
- Dissolved gaseous mercury (DGM): This includes elemental mercury (Hg), monomethylmercury (MM-Hg) and dimethylmercury (DM-Hg).

1. CLEAN LABORATORY; CLEAN BENCHES

Particles are everywhere, including dust in the air or on clothes, hair or skin. Owing to the clothes, the person who is working with the samples for trace metal analysis is the main source of contamination because this person is a particle producer. One of the most important things during sample pretreatment for trace metal analysis is to eliminate particles that can contaminate the samples or the sample containers from the laboratory environment.
The best way to eliminate most of this contamination is to work under a laminar flow box with a laminar horizontal flow (sample protection). Recommended conditions for a 'clean bench' or a 'clean lab' are class 100 (US Norm) which means that there are still about one hundred particles present per cubic foot or class 3 (DIN-Norm), which equals 3000 particles per m3 (corresponding to class 100 US Norm).

2. PREPARATIONS

Chemicals

High purity water (e.g., 'Milli-Q water', 18 M cm-1) freshly prepared, is termed 'water' in the following text.

A sub-boiling quartz still is recommended for the distillation of highly purified acids and solvents. A teflon still is recommended for the distillation of HF.

Amalgamation (filtration of oversaturated solutions with goldnet) and volatilization (bubbling with ultrapure argon) are effective methods to purify (clean) chemicals and solutions for mercury analysis.

In order to avoid contamination problems, all plastic ware, bottles and containers must be treated with acids (HCl or HNO3) for several weeks and then rinsed with water and covered in plastic bags until use.

The following procedures (Patterson and Settle, 1976) are suggested:

Laboratory ware

Store in 2M HCl (high purity) for one week, rinse with water, store in water for one week and dry under dust-free conditions (clean bench).

Samplers and bottles

Sampling devices: Fill with 1% HNO3 (high purity), store at room temperature for three weeks, and rinse with water .

Teflon/quartz bottles: Store in warm (40 C ±5 C) 1:1 diluted HCl for one week. Then rinse with water and store with 1M HNO3 (high purity) until the final use (a minimum of three weeks). Modified cleaning procedures are required for mercury. Glass containers (borosilicate, quartz) used for the collection and storage of samples for the determination of mercury are usually cleaned using an oxidizing procedure described by Sturgeon and Berman (1987). Bottles are filled with a solution of 0.1 % KMnO4, 0.1% K2S2O8 and 2.5 % HNO3 and heated for 2 hours at 80 C. The bottles are then rinsed with water and stored with 2 % HNO3 containing 0.01 % K2Cr2O7 or KMnO4 until ready for use.

Filters

Polycarbonate filters (e.g., Nuclepore) (0.4 m, 47 mm diameter) are recommended for trace metals except mercury. Store the filters in 2M HCl (high purity) for a minimum of three weeks. After rinsing with water, store for one more week in water.

For the determination of mercury, glass microfibre filters (GF/F grade, Millipore type) and teflon filters are recommended for the filtration of natural water samples. Cleaning of these filters is comparable to the procedure used for polycarbonate filters. For GF/F filters, an additional drying step has to be considered (450 C for 12-24 hr) to volatilize gaseous mercury. This procedure is described in detail by Queremais & Cossa (1997).

If trace metals in suspended particulate matter (SPM) are to be determined, filters have to be placed in precleaned plastic dishes, dried in a clean bench for two days, and stored in a desiccator until they are weighed using an electronic microbalance with antistatic properties. Each filter has to be weighed daily for several days until the weight is constant. The same procedure for drying and weighing should be applied to the filters loaded with SPM (Pohl, 1997).

3. SAMPLING AND SAMPLE HANDLING

The basis for the reliable measurement of extremely low concentrations of trace metals in sea water is a well-performed sampling to avoid contamination risk from the ship. Careful handling is recommended because copper and tin are still the main substances used in antifouling paints on ships and there is also a risk of contamination by zinc (anodes of the ship), iron or lead.
In coastal and continental shelf waters, samples are collected using 30 l teflon-coated GO-FLO (General Oceanics, close-open-close system) bottles with teflon O-rings deployed on Kevlar or on a Hostalen coated wire. Niskin bottles deployed on rosettes using standard stainless steel hydrowire are also acceptable. For surface waters, an all-teflon MERCOS-Sampler (Hydrobios) could be chosen.

PVC gloves should be worn during subsampling into the precleaned quartz or teflon bottles (teflon has an extra low content of trace metals). Subsampling should be carried out in a clean lab or a clean-lab container, if available.

Pumping of samples using peristaltic or teflon piston pumps must be carried out using precleaned silicon- or teflon-lined tubes.

In the absence of clean-lab conditions, sampling and sample handling must be carried out in a closed system, or contamination cannot be avoided.
For mercury analysis, it should be noted that the integrity during sampling and storage may be jeopardized by the addition of mercury to the sample as well as by unexpected losses owing to volatilization.

4. FILTRATION PROCEDURE

In the environmental and geochemical scientific community concerned with water analysis, it has generally been accepted that the term 'dissolved' refers to that fraction of water and its constituents which have passed through a 0.45 m membrane filter. This is an operationally defined fraction. Coastal and shelf water samples have to be filtered to eliminate particles from the water. A number of metal species pass through this filter pore size, including metals bound to colloids or clays or to humic, fulvic, amino, and fatty acids.

To prevent desorption of metal ions from particle surfaces or from biological degradation of SPM, separation between the dissolved phase and the particulate phase has to be done immediately after sampling by filtering the water through a 0.45 m polycarbonate filter. This procedure should be carried out under clean conditions (clean benches are recommended on board the ship). If metals in both the dissolved and particulate phases are to be analysed, pressure filtration with nitrogen is recommended. After filtration the filter should be rinsed with high purity isotonic solution to remove sea salt residues. Only a few millilitres are necessary because a change of pH could cause desorption of metal ions from the particles. In pumping systems, on-line filtration is possible.

5. STORAGE OF SAMPLES

To avoid wall adsorption of metal ions, 1.5 ml HNO3 or HCl (high purity) should be added per litre of seawater sample immediately after filtration for acidification to pH 1.0-1.6. The sample containers should be stored in plastic bags under controlled environmental conditions. The filters should be stored in plastic dishes at -18 C or below. Under these conditions, both water samples and SPM on filters can be stored for at least one year.

Special consideration must be given to samples destined for Hg determinations. It is necessary to add either oxidants (Cr2O72-) in addition to acidification or complexing agents (cysteine) to neutral or alkaline samples to prevent Hg losses during storage.

6. SAMPLE PRETREATMENT

Water samples

Depending on the expected concentration range (10-7-10-9 gkg-1) of trace metals (dissolved) in Baltic Sea water and because of the salt matrix interfering during the measurement process, preconcentration techniques and/or the elimination of sea salt has to be carried out prior to the analytical measurement. Detailed method information is available in the open literature (e.g., Danielsson et al., 1978; Kremling et al., 1983; and Pohl, 1994).

Filters

Different methods to analyse the material on the filter are described by Hovind and Skei (1992) and Loring and Rantala (1991). Pressure decomposition with an acid mixture (HCl, HNO3, HF) is recommended. If the silica content is high due to diatoms, the HF concentration should be increased accordingly. If the organic content increases, it is advisable to work with perchloric acid.

Depending on the digestion system used (high pressure autoclave, microwave digestion, wet ashing in an open system, or dry ashing), the completeness of the digestion is a function of temperature, time, digestion material and pressure, and has to be tested and validated in pilot studies with (certified) reference materials (see the detailed remarks in Annex B-7, Section 4.3). Digestion of samples for mercury analysis must always be carried out in a closed system to prevent losses by evaporation.

7. INSTRUMENTATION

For the analytical measurements, several analytical techniques can be used, such as GFAAS (graphite furnace atomic absorption spectrometry), electrochemical methods, ICP-MS (inductively coupled plasma-mass spectrometry), ICP-AES (inductively coupled plasma-atomic emission spectrometry), or total-reflection X-ray fluorescence (TXRF). Because of the very low mercury concentrations in sea water, the most widely used technique for mercury is the cold vapour technique (reduction of mercury with SnCl2 to elemental Hg) and preconcentration of mercury by amalgamation on a gold trap. This is followed by atomic absorption spectrometry or by atomic fluorescence spectrometry, with detection limits adequate for the purpose. In the case of anoxic (sulfur-containing waters), see Annex B-11.

8. QUALITY CONTROL

The internal quality control is described in Chapter B.5 of the Manual.

Blank

Particularly in the case of trace metal analysis, with high contamination risks at each step of the analytical work, a satisfactory blank control is necessary. Therefore, it is important to control the blank daily, for reproducibility and constancy over a longer time. The blank should include all analytical pretreatment procedures, including the addition of the same quantities of chemical substances as for the sample.

Calibration

For calibration purposes, single element standard stock solutions at a concentration of 1000 mg dm-3, purchased from a qualified manufacturer, should be available. Preparation date and concentration should be marked on the bottle. From this stock solution, a multi-element working standard solution can be prepared using dilute HCl or HNO3 as required (normally 1M acid is used).

Traceability can be ensured by the use of CRMs or participation in intercomparison exercises. The working standard should be prepared from the stock standard solution for every batch of samples and kept no longer than two weeks. Precleaned teflon containers are preferable for storage.

To evaluate effects from the matrix, the method of standard addition can be used, particularly in connection with the analytical method of voltammetric stripping. For other techniques, the method of standard addition should generally be used with care (Cardone, 1986a, 1986b).
Reference materials

Owing to problems in defining the blank, the use of a low-concentration CRM is important. Regular participation in intercomparison exercises should be considered mandatory.

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Annex III:

HELCOM (2012b). Manual for marine monitoring in the COMBINE programme. Annex B-11, Appendix 2. Technical note on the determination of persistent organic pollutants in seawater. (3.1.3)

HELCOM Manual for marine monitoring in the COMBINE programme

ANNEX B-11 APPENDIX 2: TECHNICAL ANNEX ON THE DETERMINATION OF HEAVY METALS AND PERSISTENT ORGANIC COMPOUNDS IN SEAWATER

TECHNICAL NOTE ON THE DETERMINATION OF PERSISTENT ORGANIC POLLUTANTS IN SEAWATER

1. INTRODUCTION

These guidelines concentrate on the sampling and extraction of lipophilic persistent organic pollutants from seawater and special aspects of the sampling matrix. This group of pollutants comprises the group of polycyclic aromatic hydrocarbons (PAHs) and chlorinated hydrocarbons (e.g., HCH, HCB, DDT group, chlorinated biphenyls (PCBs)).

For general aspects and the analytical determination, reference is made to the following guidelines:

• "Determination of Polycyclic Aromatic Hydrocarbons (PAHs) in Sediments: Analytical Methods", ICES ACME Report 1997;

• "Guidelines for the determination of chlorobiphenyls in sediments: Analytical methods", ICES ACME Report 1996;

• "Determination of Polycyclic Aromatic Hydrocarbons (PAH)s in Biota", ICES ACME Report 1998; and

• Annex B-14 (these Guidelines).

As the same analytical methods can be used for the determination of lipophilic pollutants in extracts of water samples as are used for extracts of sediments, it is felt that it is a useful way to unify analytical procedures to refer to these publications only.

However, it should be taken into consideration (e.g., for calibration) that the relative concentrations of the individual pollutants are generally quite different in water and sediment samples. The concentration patterns of the pollutants are mainly influenced by their polarity which can be expressed by their octanol/water coefficient (log Kow; Kow = Concentration in octanol phase / Concentration in aqueous phase). Thus, in water samples the more hydrophilic compounds with log Kow values of 3 to 4 predominate (e.g., 2- and 3-ring aromatics and HCH isomers), while in sediments and biota the pollutants with log Kow values >5 are enriched (4- to 6-ring aromatics, DDT group, PCBs).

These guidelines provide advice on lipophilic persistent organic pollutant (POPs) analyses in total seawater with a log KOW > 3. The analysis of POPs generally includes:

• sampling and extraction of the water;
• clean-up; and
• analytical determination

The extraction of the POPs simultaneously enables an enrichment of the analytes. Because of the very low concentration range of 10 pg l−1 to 10 ng l−1, the enrichment of the contaminants is a very important step in the procedure. Extraction and enrichment can be done by solid phase extraction (SPE) or liquid-liquid extraction (LLE).

Determination depends on the chemical structure of the compounds. PAHs can be determined by high performance liquid chromatography (HPLC) with fluorescence detection or gas chromatographic (GC) separation with flame ionization (FID) or mass spectrometric (MS) detection (Fetzer and Vo-Dinh, 1989; Wise et al., 1995). Chlorinated hydrocarbons are generally analysed by gas chromatographic (GC) separation with electron capture detectors (ECD) or mass spectrometric (MS) detection.

All steps of the procedure are susceptible to insufficient recovery and/or contamination. Therefore, regular quality control procedures must be applied to check the performance of the whole method. These guidelines are intended to encourage and assist analytical chemists to critically reconsider their methods and to improve their procedures and/or the associated quality control measures, where necessary.

These guidelines are not intended as a complete laboratory manual. If necessary, guidance should be sought from specialized laboratories. Whichever procedure is adopted, each laboratory must demonstrate the validity of each step of its procedure. In addition, the use of a second (and different) method, carried out concurrently to the routine procedure, is recommended for validation. The participation in analytical proficiency tests is highly recommended.

2. SAMPLING AND STORAGE

Plastic materials must not be used for sampling and storage owing to possible adsorption on the container material or contamination. Especially the very lipophilic compounds (4- to 6-ring aromatic hydrocarbons, DDT, PCBs) tend to adsorb on every surface. Therefore, the seawater samples should not be stored longer than 2 h and should not be transferred into other containers before extraction. It is highly recommended to extract the water sample as soon as possible after sampling and to use as little manipulation as possible. It is recommended that sampling and extraction should be done in the same device. Extracts in organic solvents are less susceptible to adsorption onto surfaces.

3. BLANKS AND CONTAMINATION

In many cases, the procedural detection limit is determined by the blank value. In order to keep the blank value as low as possible, the compounds to be analysed or other interfering compounds should be removed from all glassware, solvents, chemicals, adsorption materials, etc., that are used in the analysis. The following procedures should be used:

- Glassware should be thoroughly washed with detergents and rinsed with an organic solvent prior to use. Further cleaning of the glassware, other than calibrated instruments, can be carried out by heating at temperatures > 250 °C.

- All solvents should be checked for impurities by concentrating the amount normally used to 10 % of the normal end volume. This concentrate is then analysed in the same way as a sample by HPLC or GC and should not contain significant amounts of the compounds to be analysed or other interfering compounds.

- All chemicals and adsorption materials should be checked for impurities and purified (e.g., by heating or extraction), if necessary. Soxhlet thimbles should be pre-extracted. Glassfiber thimbles are preferred over paper thimbles. Alternatively, full glass Soxhlet thimbles, with a G1 glass filter at the bottom, can be used. The storage of these supercleaned materials for a long period is not recommended, as laboratory air can contain PAHs that will be adsorbed by these materials. Blank values occurring despite all the above-mentioned precautions may be due to contamination from

the air. The most volatile compounds will usually show the highest blanks (Gremm and Frimmel, 1990).

As the concentrations of the PAHs and chlorinated hydrocarbons in seawater are very low, possible blank and contamination problems might be even more difficult to control than with sediment samples. Therefore, it is recommended to rewash all equipment (vials, pipettes, glass bottles) with solvent just before use. If possible, critical steps should be done in a clean bench.

The more volatile compounds (especially naphthalene and phenanthrene) show the largest blank problems.

4. PRE-TREATMENT

For the extraction of whole water samples, no pre-treatment is necessary.
If the suspended particulate material (SPM) will be analysed separately from the solute phase, a phase separation has to be done. Because of the necessary additional manipulation step, this is a difficult operation which affords a number of additional quality control procedures (adsorption losses, contamination problems). There are two possible ways for phase separation: filtration and centrifugation.

Filtration is done by GF/F glass fibre filters. As flat-bed filters have a very limited capacity, the use of coiled glass fibre filters is recommended for volumes larger than 10 l and water samples with high amounts of suspended matter. A pump is necessary to force the water through the filter. Centrifugation needs a high volume centrifuge which must be operable onboard a ship. Such centrifuges with a throughput of 1 m^3 $h^{-1}$and more are commercially available and used for sampling SPM; however, they are expensive and generally not a standard equipment. For centrifugation, blanks and adsorption problems have to be controlled as well as the separation efficiency.

The sampled SPM is analysed like a sediment. The solute phase is analysed like the whole water sample.

Validation of the phase separation procedures is very difficult; thus, it might be wise to analyse the whole water sample for monitoring purposes and to determine separately only the amount of SPM in the water for reference or normalization purposes.

5. EXTRACTION

The volume of the water sample is the most important parameter which influences the limit of determination of the method. As POP concentrations down to 10 pg l^{-1} and less are observed in seawater, large water volumes of 10 l to 100 l have to be sampled and extracted. Large volumes are required not only to obtain a sufficiently high detector signal, but also to discriminate from blank problems.

Principally, there are two different extraction principles in current use: solid phase extraction (SPE) and liquid-liquid extraction (LLE). Unfortunately, the two procedures do not always yield comparable results, as the physical extraction principles are quite different (Sturm et al., 1998, Gomez-Belinchon et al., 1988).

SPE has the advantage of being able to extract very large water volumes (up to 1000 l) and to incorporate a phase separation to obtain separate samples for SPM and the solute phase. The

drawbacks of the method are a longer sampling time demand, a more complex instrumentation, and problems with validation and control of the extraction efficiency.

LLE has the advantage that it can be easily validated and controlled, as internal standards can be added before extraction. Also, standard addition techniques can be used for accuracy testing. As LLE is a classical extraction technique, a great deal of experience is available and the robustness of the principle is proven. The limitation in sample volume is only relative, as techniques have been described for sampling 10 l and 100 l on a routine basis (Gaul and Ziebarth, 1993; Theobald et al., 1990). It has been shown that a sampling volume of 100 l is sufficient for nearly all monitoring tasks. Because of the robustness of the method, there is a preference LLE for routine monitoring purposes for all lipophilic organic contaminants.

5.1 Solid phase extraction

The extraction device consists of a filter holder, an adsorption column filled with an adsorbing material (e.g., XAD resin, C18 modified silica gel), a pump which forces the water sample through the column, a flow meter, an electronic control unit, and a power supply. Sampling can be done either by deploying the whole extraction device into the water (in situ pumping) or by pumping the water with a separate pump onboard a ship and then through the extraction device. A suitable in situ system is described in detail in Patrick et al. (1996). After sampling, the columns are stored at 4 °C and the filters at –20 °C.

The adsorption column is eluted with an organic solvent (acetone or acetonitril). Prior to the extraction, internal standards are added to the solvent. The extract obtained is pre-cleaned and analysed.

Analytical procedures for the use of XAD-2 adsorption resins are published by the IOC (1993), Ehrhardt (1987), and Bruhn and McLachlan (2001).

Although the SPE technique has many advantages, one has to be aware of some problems. Especially for large volume sampling, validation of the method is extremely difficult and has not yet been achieved. Some publications have shown that the extraction efficiency is dependent on, e.g., the amount and kind of humic substances which can complex lipophilic compounds (Johnson et al., 1991; Kulovaara, 1993; Sturm et al., 1998).

5.2 Liquid-liquid extraction

The decision to sample 10 l, 20 l, or 100 l of water depends on the anticipated concentrations of the compounds to be analysed in natural samples . For remote sea areas with expected concentration of 10 pg l^{-1} or less, a volume of 100 l is recommended. The technique and principle are identical for all volumes, only the sampling bottle and the equipment are different. Details of the sampling and extraction techniques are described in Gaul and Ziebarth (1993) for the 10 l sampler and in Theobald et al. (1990) for the 100 l sampler.

The all-glass bottle sampler fixed in a stainless steel cage is lowered by a hydrographic wire down to the sampling depth and opened under water. After filling, the sampler is brought on deck of the ship and immediately extracted with a non-polar solvent such as pentane or hexane. Prior to extraction, a solution with appropriate internal standards (e.g., deuterated PAHs, e-HCH, PCB 185) is added to the water sample. After phase separation, the organic extract is dried with Na_2SO_4 and carefully concentrated to about 1 ml in a rotary evaporator. Further evaporation is done under a gentle stream of nitrogen.

Extreme care has to be taken to ovoid contamination during sampling, extraction, and work up. Blank samples must be taken in every sampling campaign; this can be done, e.g., by rinsing the cleaned sampling bottle with the extraction solvent and treating this extract like a normal sample. The sampling bottle must be cleaned with detergent, water, and organic solvents (acetone and hexane or pentane) before use. After using in open sea areas, it can be of advantage not to perform the whole cleaning/washing procedure but just to use the sampler directly after emptying the glass bottle from the extracted previous water sample.

Extracts should be stored in the refrigerator and in the dark.

6. CLEAN-UP

Interferences from matrix compounds in seawater samples are generally smaller than in sediment or biota samples. Nevertheless, the crude extracts require a clean-up before chromatographic separation and determination can be done. The clean-up is dependent on the compounds to be analysed, the sample, the determination method used, and the concentration range to be analysed. For all GC methods, it is essential to remove polar and non-volatile compounds in order to protect the GC column from rapid destruction. A detection system with low selectivity (eg., GC-FID) needs a far better clean-up than a detector with a high selectivity such GC-MS or even GC-MS/MS. HPLC with fluorescence detection (for PAH analyses) has a relative high selectivity but the method will fail if petrogenic aromatic compounds (from an oil spill) are present in the sample. GC-ECD (for chlorinated compounds) has a high selectivity but some interferences (e.g., phthalate esters) may disturb the detection; therefore, for GC-ECD a good clean-up is necessary as well.

A clean-up procedure for this is presented here that uses short silica gel chromatography columns that can be applied with any determination technique: HPLC, GC or GC-MS. The method is simple and is sufficient in most cases of PAH and chlorinated hydrocarbon determinations in seawater (ICES, 1996, 1997, 1999).

A 3 ml glass column with glass fibre frit (commercially available for SPE) is filled with 500 mg silica gel (dried for 2 h at 200° C) and subsequently washed with 30 ml CH2Cl2 and 30 ml hexane. The hexane sample extract (concentrated to 500 µl) is applied on top of the column and eluted with 5 ml CH2Cl2/hexane (15/85 v/v) and then with 5 ml of acetone. Fraction 1 contains all lipophilic compounds of interest (PAHs and all chlorinated hydrocarbons (from HCB to HCH)); this fraction can be used for GC-MS determination after concentration to 50–300 µl. If the water sample has been extremely rich in biological material (algae) or if detection limits far below 10 pg l−1 are requested, additional clean-up (HPLC, GPC) might become necessary.

7. CROMATOGRAPHIC DETERMINATION

Details for the chromatographic determinations are comprehensively described in the 1996 ACME report (ICES, 1996) for chlorobiphenyls in sediments (GC-ECD and GC-MS), the 1997 ACME report (ICES, 1997) for PAHs in sediments (HPLC-Fluorescence detection, GC-FID and GC-MS), and the 1998 ACME report (ICES, 1999) for PAHs in biota (HPLC and GC-MS).

As the cleaned extracts from the seawater samples can be analysed in the same way as the extracts from sediments and biota, the above guidelines can be used. When a GC-MS system can be used, all compounds can be determined in one single GC analysis; if not, the samples have to be analysed separately for PAHs (HPLC-F, GC-FID) and chlorinated hydrocarbons (GC-ECD).

7.1 Gas chromatography-mass spectrometry

As GC-MS has the advantage of being both very selective and quite universal, it is strongly recommended to use GC-MS as the determination method. It especially has the advantage that both PAHs and chlorinated hydrocarbons can be determined in one single analysis. This is not possible with any of the other techniques.

Because of the sensitivity required, the mass spectrometric detector must be operated in the selected ion mode (SIM). By this, absolute sensitivities in the range of 1 pg to 10 pg can be achieved for most compounds. Ion-trap instruments can be operated in full-scan mode and are in principle as sensitive as quadrupole detectors; however, with real samples and matrix underground they can lose considerably sensitivity.

With GC-MS, detection limits of 5–30 pg l^{-1} can be reached with water sample volumes of 10 l to 100 l. In most cases, it is not the absolute signal strength of the detector which limits the detection; therefore, the injection of a larger aliquot of the analysis solution would not improve it. For some compounds, blank values are the limiting parameter (especially naphthalene and phenanthrene and, to a lesser extent, other PAHs); for this, only a larger sample volume can improve the detection limits. Many other compounds do not exhibit blank problems, if appropriate care is applied; for these, matrix noise often limits the detection. For such situations, only a better clean-up (e.g., HPLC, GPC) or a more specific detection method (GC-NCI-MS or GC-MS/MS) will improve the detection limit. Negative chemical ionization (NCI) mass spectrometric detection can be used for highly chlorinated compounds (e.g., HCB, PCBs with five or more Cl atoms, HCH) and shows extremely high sensitivity and selectivity for these compounds. More universally applicable is tandem mass spectrometry (MS/MS), which yields a similar absolute sensitivity as normal MS but much higher selectivity. Some MS/MS transitions for the detection of selected chlorinated hydrocarbons are listed in Table 1 in Appendix 2 to Annex B-13: Technical note on the determination of polycyclic aromatic hydrocarbons in biota, from the full "Guidelines".

7.2 Quantification

A multilevel calibration with at least five concentration levels is recommended. The response of the FID detector is linear. For UV and fluorescence detection, the linear range is also large. The working range should be linear and must be covered by a calibration curve. Since the mass spectrometric detector often has no linear response curve, the use of stable deuterated isotopes is a prerequisite. Furthermore, the response of PAHs in standard solutions is often much lower than in sample extracts. Only a combination of different techniques, e.g., the use of internal standards and standard addition, might give reliable quantitative results.

The calibration curve can be checked by recalculating the standards as if they were samples and comparing these results with the nominal values. Deviations from the nominal values should not exceed 5%.

When chromatograms are processed using automated integrators, the baseline is not always set correctly, and always needs visual inspection. Because the separation of the peaks is often incomplete in HPLC analysis, the use of peak heights is recommended for quantification. In case of GC techniques, either peak heights or peak areas can be used.

Prior to running a series of samples and standards, the GC or HPLC systems should be equilibrated by injecting at least one sample extract, the data from which should be ignored. In addition, standards used for multilevel calibration should be regularly distributed over the sample series so matrix- and non-matrix-containing injections alternate. A sample series should include:

• a procedural blank,
• a laboratory reference material,
• at least five standards,
• one standard that has been treated similarly to the samples (recovery determination).

The limit of determination should depend on the purpose of the investigation. A limit of 2 ng g^{-1} (dry weight) or better should be attained for single compounds. The method for calculating the limit of determination should reflect QUASIMEME advice (Topping et al., 1992). The limit of determination that can be achieved depends on the blank, the sample matrix, concentrations of interfering compounds, and the volume of water taken for analysis. The typical concentration ranges of PAHs and other POPs in seawater can be found in HELCOM assessments (HELCOM, 2003a, 2003b).

8. QUALITY ASSURANCE

A number of measures should be taken to ensure a sufficient quality of the analysis. Five main areas can be identified:

1. extraction efficiency and clean-up;
2. calibrant and calibration;
3. system performance;
4. long-term stability; and
5. internal standards.

8.1 Extraction efficiency and clean-up

A check on extraction efficiency and clean-up can be performed by analysing a reference material (Annex B-7). To determine the recovery rates of the clean-up and concentration steps, it is recommended to pass a standard solution through the entire procedure. Additionally, at least one internal standard should be added to each sample before extraction, to check for recovery during the analytical procedures. If major losses have occurred, then the results should not be reported. CB29 is suggested as a recovery standard because, owing to its high volatility, losses due to evaporation are easily detected. CB29 elutes relatively late from alumina and silica columns. Small peaks that may be present in the gas chromatogram at the retention time of CB29 do not hinder the use of this CB because the recovery standard only indicates major errors in extraction or clean-up. In case of GC/MS, labelled CBs can be used as recovery standards. This allows correction for recovery, provided that each chlorination stage is represented.

8.2 Calibrant and calibration

PAH determinations should preferably be carried out using calibration solutions prepared from certified crystalline PAHs. However, the laboratory should have the appropriate equipment and expertise to handle these hazardous crystalline substances. Alternatively, certified PAH solutions, preferably from two different suppliers, can be used. Two independent stock solutions should always be prepared simultaneously to allow cross-checks to be made. Calibration solutions should be stored in ampoules in a cool, dark place. Weight loss during storage should be recorded for all standards.

CB determinations should always be carried out using calibration solutions prepared from crystalline CBs. Preferably, certified CBs should be used. Two independent stock solutions of different concentrations should always be prepared simultaneously to allow a cross-check to be made.

Calibration solutions should preferably be stored in a cool, dark place. For all containers with standards, the weight loss during storage should be recorded.
After clean-up and before GC analysis, both in PAH and CB analysis, an additional internal standard is added for volume correction. Internal standards should be added in a fixed volume or weighted to all standards and samples.

8.3 System performance

The performance of the HPLC or GC system can be monitored by regularly checking the resolution of two closely eluting PAHs or CBs. A decrease in resolution indicates deteriorating HPLC or GC conditions. The signal-to-noise ratio of a low concentration standard yields information on the condition of the detector. For example, a dirty MS-source can be recognized by the presence of a higher background signal, together with a reduced signal-to-noise ratio. Additionally, the peak can be affected.

8.4 Long-term stability
One laboratory reference sample should be included in each series of samples. A quality control chart should be recorded for selected PAHs, e.g., fluoranthene (stable results), pyrene (sensitive to quenching), benzo[a]pyrene (sensitive to light), or, correspondingly, for selected CBs. If the warning limits are exceeded, the method should be checked for possible errors. When alarm limits are exceeded, the results obtained should not be reported. A certified reference material (CRM) should be analysed at least once a year, when available, and each time the procedure is changed. Each laboratory analysing PAHs and CBs in water should participate in interlaboratory analytical performance tests on a regular basis.

8.5 Internal standards

Internal standards should be added to all standards and samples either in a fixed volume or by weight. The PAH internal standards should preferably be non-natural PAHs which are not found in water and do not co-elute with the target PAHs; several predeuterated PAHs have proved to be suitable for GC/MS as well as for HPLC analysis. For example, for GC/MS it is recommended to add four internal standards representing different ring-sizes of PAHs.

The following compounds can be used (Wise et al., 1995):

• for HPLC analysis: phenanthrene-d10, fluoranthene-d10, perylene-d12, 6-methyl-chrysene;
• for GC/MS analysis: naphthalene-d8, phenanthrene-d10, chrysene-d12, perylene-d12;
• for GC/FID analysis: 1-butylpropylene, m-tetraphenyl.

Similarly the ideal internal standard for PCBs is a compound which is not found in the samples and does not co-elute with other CBs, e.g., CBs 29, 112, 155, 198 or all 2,4,6-substituted CB congeners. Alternatively, 1,2,3,4-tetrachloronaphthalene can be used.

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